

Sentinel-2 high-resolution data for river discharge monitoring

Paolo Filippucci (1, 3, 4), Luca Brocca (1), Stefania Bonafoni (2), Carla Saltalippi (3), Wolfgang Wagner (4),
Angelica Tarpanelli (1)

1. National Research Council, Research Institute for Geo-Hydrological Protection, Perugia, Italy
(paolo.filippucci@irpi.cnr.it)
2. Università degli Studi di Perugia, Department of Engineering, Perugia, Italy
3. Università degli Studi di Perugia, Department of Civil and Environmental Engineering, Perugia, Italy
4. TUWien (Technische Universität Wien), Department of Geodesy and Geoinformation, Wien, Austria

Abstract

River monitoring is an open issue due to many intrinsic problems of the ground monitoring network. Over the last few decades, the role of satellite sensors in river discharge estimation is significantly increased thanks to the strong growth in technologies and applications. Focusing on daily river discharge measurements, a non-linear regression model has been used to link the near-infrared (NIR) reflectance ratio between a dry and a wet pixel around the section of a river to the ground measurements of river discharge. The use of medium-resolution satellite data, such as those from MODIS sensors, enables to monitor high and low flows in medium-sized catchments (<100'000 km²), thanks to satellite frequent revisit time and wide spatial coverage. However, such sensors are not suitable to provide information for medium-narrow rivers (< 250 m wide), nor to study river features and patterns that are averaged within a single pixel. Here, we investigated the use of Sentinel-2 NIR reflectances to support the hypothesis that a higher spatial resolution, i.e. 10 m, is able to better identify the wet pixels, more related to the river dynamics, with obvious advantages for river discharge estimation compared to the medium resolution sensors (e.g., MODIS at 250 m). Moreover, it also allows both a finer distinction between vegetation, soil and water and the characterization of water turbidity

24 in the river area. A new formulation enriched by the sediment component is proposed together with a first
25 step towards an uncalibrated procedure to select the wet pixels. Google Earth Engine (GEE) platform has
26 been employed for the data analysis, allowing to avoid the download of big amounts of data, fostering the
27 reproducibility of the analysis in different locations. The accuracy of the river discharges derived from
28 Sentinel-2 reflectances is evaluated against the in-situ observations from selected gauging stations along two
29 Italian rivers, Po and Tiber. The results confirm the good performances obtained with high-resolution images
30 over the Po River, with average Nash-Sutcliffe efficiency ranging between 0.39 and 0.56 for the different
31 configurations adopted. Relatively worse results were obtained over the Tiber River where the Nash-Sutcliffe
32 efficiency ranged between 0.2 and 0.61, due to an issue on the registration of Sentinel-2 images.

33 **Keywords:** NIR reflectance, river discharge, Sentinel-2, sediment transport, Google Earth
34 Engine, MODIS

35 **1. Introduction**

36 River discharge is one of the Essential Climate Variables, which contribute most to the characterization of
37 Earth's climate and its changes, as recognized by the Global Climate Observing Systems. On long time scales,
38 the knowledge of river discharge values is important to identify potential effects of climate change over
39 lands and ocean, since their changes could affect both ocean salinity and thermohaline circulation (Nohara
40 et al. 2006; Piecuch et al., 2018; Ahmed et al., 2020). In shorter terms, it is a variable fundamental for the
41 water cycle closure and for numerous applications of water management and related services, including
42 flood protection (Hannah et al., 2011).

43 Currently, in-situ measurements are the most cost-effective and reliable option for river discharge
44 monitoring (Fekete et al., 2015). However, there are still many rivers where discharge estimates either do
45 not exist or are not promptly available, sometimes due to economic or political reasons (Fekete and
46 Vörösmarty, 2007). This is particularly true in developing countries, even if the number of working gauging
47 stations is declining worldwide (Crochemore et al., 2020; Vörösmarty et al., 2001).

48 Earth Observation (EO) data are currently a valuable source of information capable to support ground-
49 based networks. In recent years, several approaches have been developed to infer river discharge
50 information from satellite data. These techniques can be divided into those that need to be calibrated
51 against observed data and those that do not require calibration and can therefore be applied even in
52 ungauged basins (Gleason and Durand, 2020). The latter approaches use as input only remote sensing data
53 or other global data that are independent of specific calibration. River discharge can thus be obtained using
54 different paradigms. Some studies focus on developing hydrologic models to parse the components of the
55 hydrologic cycle (precipitation, evapotranspiration, terrestrial storage) obtained by remote sensing and
56 derive river discharge from the associated water excess (e.g. Emery et al., 2018; Lopez et al., 2017; Parr et
57 al., 2015). An other paradigm consists of exploiting the laws of fluid mechanic to develop hydraulic models
58 that can infer river discharge from hydraulic variables, such as elevation, slope, and width (e.g. Larnier et
59 al., 2020; Oubanas et al., 2018). Finally, Mass conserved Flow law Inversions (McFlI, Gleason et al., 2017)
60 applied to flow equations like the Manning's equation (e.g. Durand et al., 2014; Garambois and Monnier,
61 2015) or hydraulic geometric power laws (e.g. Feng et al., 2019; Gleason and Smith, 2014) can be exploited
62 to infer river discharge in ungauged rivers. In all of these cases, the obtained river discharge is typically
63 characterized by low accuracy, because the problem of uncalibrated river discharge estimation is basically
64 ill posed: critical variables for accurately measuring river discharge, such as bathymetry or friction, cannot
65 be measured by remote sensing (Gleason and Durand, 2020). Nonetheless, uncalibrated models can
66 improve the scarce information available on ungauged rivers, thus even inaccurate models are useful.
67 Alternatively, in-situ surveys and Unmanned Aerial Vehicles (UAVs) can be exploited to obtain additional
68 information on ungauged river systems (Yang et al., 2019), also coupled with satellite data (Lou et al., 2020;
69 Wufu et al., 2021), thus improving the obtained performance.

70 Calibrated approaches are usually more reliable, being adjusted using in-situ data or calibrated hydrologic
71 models. They are mainly useful for extending the knowledge of river discharge in space and time. The
72 majority of EO satellite-based techniques to retrieve river discharge belong to this category, according to
73 different paradigms. Calibrated hydrologic models can be used to improve the knowledge of river discharge

74 as, e.g., in Dziubanski et al. (2016), where a hydrologic model with a remote sensing-based snow routine is
75 used to improve river discharge estimation for seven watersheds in the Upper Mississippi River basin. This
76 is just one example of a rich literature where one or more satellite products are used in
77 calibration/assimilation to retrieve river discharge information not only at regional (Jodar et al., 2018; Syed
78 et al., 2005; Wulf et al., 2016) but also at global scale (Chandanpurker et al., 2017; Lin et al., 2019; Zhang et
79 al., 2016). Hydraulic models are also widely used (e.g., Bjerklie et al., 2003; 2005; De Frasson et al., 2019),
80 sometimes coupled with hydrological modeling (e.g., Neal et al., 2009). Finally, many calibrated
81 methodologies for estimating river discharge are based on the concept of the space-based rating curves. In
82 these approaches, satellite measurements of river level, river width or other empirical fluvial geomorphic
83 phenomena are coupled with observed river discharge data to calibrate empirical relationships that can be
84 used to effectively estimate river discharge from remote sensing data. In this framework, radar altimetry
85 and optical sensor measurements are the most promising source of data, although microwave data are
86 sometimes used as well (Brakenridge et al., 2007; Huo et al., 2021). Altimetric data are widely used because
87 of the direct relation between stage and river discharge (Abdalla et al., 2021; Belloni et al., 2021; Paris et
88 al., 2016; Tourian et al., 2013; Zakharova et al., 2020), but the spatial-temporal sampling of the altimetry
89 missions is currently a limitation, leading to the need of applying different strategies to densify the data
90 (Tourian et al., 2016; 2017; Boergens et al., 2017; Schwatke et al., 2015). As already demonstrated, if the
91 altimetry is combined with other satellite information, the estimation of river flow is improved (see
92 Tarpanelli et al., 2021 for a review). In this respect, satellite optical sensors play an important role together
93 with radar altimetry for improving the river monitoring. Specifically, water and soil are rather
94 distinguishable from optical data (Yang et al., 2020). River width measurements can therefore be extracted
95 from cloud-free images by exploiting these differences. River discharge measurements can be then
96 obtained from the river width by using empiric formulas, rating curves or models (Elmi et al., 2021; Pavelsky
97 et al., 2014), also in combination with altimetric data (Sichangi et al., 2016). Notwithstanding this, the
98 measurement is possible only in free clouds days (reducing the temporal availability of the obtained datum)

99 and there are often misclassifications of water pixels due to cloud shadows or in areas characterized by
100 steep terrain or dense urban environments (Mueller et al., 2016).

101 Focusing on daily measurements of river discharge, Tarpanelli et al. (2013; 2017; 2020) used passive MODIS
102 remote sensing data to develop a linear regression approach first introduced by Brakenridge et al. (2007),
103 which is effectively able to monitor high and low flow in medium basins (<100'000 km²). The approach
104 (hereafter named *CM*) is based on the different behavior in the NIR region between a Calibration (*C*) and a
105 Measurement (*M*) pixels chosen over land and water, respectively. Indeed, in a bare soil or urban area, the
106 *C* pixel has a reflectance behavior stable in time, while a pixel around the bank of a river, often inundated,
107 has generally lower reflectance values that are inversely proportional to river discharge (Tarpanelli et al.,
108 2013; 2017). Both pixels are affected by noises due to different sources, such as the atmosphere or the
109 acquiring process, but the calculation of their ratio allows to account for them and therefore it is well
110 correlated with the in-situ measurements of river discharge. Notwithstanding this, the presence of
111 sediments in the river induces an increase in the water reflectance, which corresponds to an
112 underestimation of the river discharge assessed through *CM* approach. The analysis of this phenomenon
113 for medium-narrow rivers (100 - 300 m) can be facilitated by the use of high-resolution multispectral
114 sensors to better represent the dynamic of such rivers. With respect to the medium resolution images used
115 so far (i.e., MODIS, MERIS and OLCI), data with higher spatial resolution coming from the new launched
116 Sentinel-2 Mission of the Copernicus European Earth Observation programme (10, 20 and 60 m spatial
117 resolution), have opened new possibilities to investigate small targets. The *CM* approach has been already
118 tested with the Sentinel-2 high-resolution data harmonized with Landsat-8 images in the Murray-Darling
119 basin in Australia, with encouraging results (Shi et al., 2020) but for a short period of time (only one water
120 year was considered) and by considering only a single pixel *M* of 30x30 m. Moreover, Shi et al. indicated as
121 a potential issue the high presence of sediment in the river water during floods events, which introduce a
122 source of noise in the *M* signal and consequently in the obtained river discharge.

123 Based on these premises, the first objective of this paper was to understand the benefits of high-resolution
124 imagery by Sentinel-2 by extending the procedure to small targets and testing its limitations. For this

125 purpose, two Italian rivers were selected as study areas: the Po River in northern Italy and the Tiber River
126 in central Italy. The first one was selected due to its medium-narrow size and the availability of multiple
127 stations with a dense record of hydraulic variables, while the second one was selected for its width of about
128 50-80 m. In particular, a total of seven stations were selected: five stations along the Po River (Piacenza,
129 Cremona, Borgoforte, Sermide and Pontelagoscuro) and two stations along the Tiber River (Montemolino
130 and Ponte Felice). In order to check the effective advances of the Sentinel-2 spatial resolution with respect
131 to medium resolution, the same analyses were carried out also using MODIS product as input. As second
132 objective, we attempted to improve the estimation of river discharge by implementing modifications at the
133 original *CM* approach related to the impact of the solid transport. Finally, we moved a first step towards an
134 uncalibrated procedure by selecting the sensitive pixels to the river discharge variations through an analysis
135 of their correlations against land and water pixels.

136 In order to foster the analysis to large time scales and in several areas, the code was implemented in the
137 Google Earth Engine, GEE, platform (Gorelick et al., 2017), reducing by far the computational time and
138 storage capacities required to download images and apply the methodology.

139 The paper is structured as follows: the study areas with the satellite and ground datasets are introduced in
140 Section 2, together with the cloud platform GEE. In Section 3, the *CM* approach and the proposed
141 modifications are described. A sensitivity analysis of the threshold adopted within the approaches is carried
142 out in Section 4, together with the results of the proposed models application on the sections of the two
143 rivers. Finally, in Section 5 the conclusions of the analysis are summarized, including limitations and future
144 developments of the procedure.

145 **2. Study area & Data**

146 **2.1 Study area**

147 The analysis is focused on seven sites located in Italy: five along the Po River at the stations of Piacenza
148 (45.11 °N, 9.64 °E), Cremona (45.13 °N, 9.97 °E), Borgoforte (45.04 °N, 10.81 °E), Sermide (44.99 °N, 11.33

149 °E) and Pontelagoscuro (44.89 °N, 11.60 °E) and the two along the Tiber River at the stations of
150 Montemolino (42.80 °N, 12.40 °E) and Ponte Felice (41.95 °N, 12.50°E). These sites were selected on the
151 basis of the availability of in-situ flow measurements during the analyzed periods (2015-2020), the different
152 regime of flow and the width of the river, which is medium-small.

153 The Po River is located in Northern Italy and it has a drainage area of 74'000 km²: it is the longest Italian
154 river and the fifth European river for river discharge. During the analyzed period the average discharge
155 value at the five stations varies between 818 and 1312 m³/s, with the minimum and the maximum values
156 being equal to 242 and 8950 m³/s, respectively. The maximum cross-section width at the bankfull discharge
157 is around 500 m (Domeneghetti et al., 2018), during the seasonal peaks in spring (May-June) and autumns
158 (November). The river discharge data are available online from the website of the Regional Agency for
159 Environmental Protection (www.arpae.it/documenti.asp).

160 The Tiber River in central Italy has a drainage area of 17'375 km² and its average river discharge is much
161 smaller than the one of the Po River: during the analyzed period, it is around 30.4 m³/s at Montemolino
162 and 117.2 m³/s at Ponte Felice, with the minimum and maximum extremes being equal to 1.6 and 804.2
163 m³/s, respectively, for Montemolino station and equal to 49.4 and 843.3 m³/s, respectively, for Ponte Felice
164 station. The river width varies between about 50 and 80 meters, surrounded by riparian vegetation. The
165 data of Montemolino are provided by the Department of Environment, Planning and Infrastructure of the
166 Umbria Region, which is in charge for the collection of hydro meteorological data in the Umbria Region,
167 while those of Ponte Felice are made available from the Department of Environment, Planning and
168 Infrastructure of the Lazio Region.

169 **2.2 Satellite Near Infrared data**

170 In order to assess the benefits deriving from the use of high spatial resolution imagery in estimating river
171 discharge, two datasets of NIR data characterized by high and medium spatial resolution were selected
172 from Sentinel-2 and MODIS sensors, respectively.

173 The Sentinel-2 mission is composed of two satellites, Sentinel-2A (launched on the 23 June 2015) and
174 Sentinel-2B (launched on 7 March 2017). The satellites are placed in the same sun-synchronous orbit,
175 phased 180° to each other. This strategy permits to increase the revisit time of Sentinel-2 from 10 days
176 (using one satellite) to 5 days (using both of them), with increased revisit frequency in the areas where
177 adjacent orbits overlap. Each satellite carries a MultiSpectral Instrument (MSI) sensor, capable of sensing
178 electromagnetic radiations in 13 spectral bands, from visible to short-wave infrared. In this analysis, data
179 from band 8 (central wavelength ~833 nm, bandwidth=106 nm) are selected, characterized by a spatial
180 resolution of 10 meters. Two products are available: Level-1C product with orthorectified TOA (Top-Of-
181 Atmosphere) reflectance and Level-2A product with surface reflectance. Generally, the atmospheric
182 correction applied to produce the Level-2 product over land is different from the one applied over
183 ocean/sea. The river is a fairly complex and mixed environment to be managed. In the absence of detailed
184 studies on this, and because the effects of the atmosphere in clear-sky conditions can be assumed negligible
185 when computing the ratio of neighboring pixels in the same area (as used in this study, see Section 3), we
186 performed the analysis by using Level-1C product.

187 MODIS is a multispectral sensor on-board the TERRA and AQUA satellites, launched on the 18 December
188 1999 and the 4 May 2002, respectively, and capable of acquiring Earth's surface data at 36 different spectral
189 bands with spatial resolution ranging from 250 to 1000 m at daily time scale. In the present study, the
190 reflectance in band 2 (Near-Infrared, central wavelength ~859 nm, bandwidth=35 nm) of level-2 product
191 MYD09GQ version 6 at 250 m of spatial resolution were considered. For cloud masking, the "state_1km"
192 band of MYD09GA version 6 product was selected. Both the products come from the MODIS sensor onboard
193 AQUA satellite.

194 The dimension of each Sentinel-2 image (~600 MB) and the number of images to be downloaded for a
195 temporal analysis, make the process of collecting and reading the data not easy and computationally
196 challenging. In order to show the method and illustrate the new algorithm, we referred to the Sentinel-2
197 data Level-1C for bands 8 dataset, provided by the Meteorological Environmental Earth Observation (MEEEO
198 – www.meeo.it) company, together with the relative cloud mask probability product. This collection

199 includes a subset of Pontelagoscuro area in the period January 2018 – September 2019 and it is used to
200 generate the figures shown in the Section 3. However, Google Earth Engine platform was used for the
201 analysis relative to the period 2015-2020 over the seven stations for both the MODIS and Sentinel-2
202 analysis, in order to lighten the computational load: the results and conclusions at Sections 4 and 5 are
203 referring to the cloud data.

204 **2.3 Google Earth Engine cloud platform**

205 Following the algorithm definition, the analysis at the study sites was carried out through Google Earth
206 Engine, GEE (Gorelick et al., 2017), a cloud computing platform for processing satellite imagery and many
207 kinds of geospatial and observation data. GEE comprehends a catalogue of satellite imagery and geospatial
208 datasets of over twenty petabytes at planetary scale, together with the analysis capabilities needed to
209 process them. The data are freely available for scientist, researcher and developers, who can access and
210 process them through the Earth Engine API using both Javascript or Python. The development of Web-
211 Application is also possible from the code editor interface. In this study, we developed a code to read,
212 extract and analyze NIR data at the considered gauging stations.

213 **3. Method**

214 In this section, the original CM approach, as proposed by Tarpanelli et al. (2013) and subsequently modified
215 by Tarpanelli et al. (2017), is described and adapted to the Sentinel-2 data. In particular, the higher spatial
216 resolution with respect to previous studies allows to highlight two aspects that have been neglected so far:
217 1) the position of the M pixel, which might be different for low and high flows, and 2) the influence of
218 suspended sediments on the C/M ratio and, thus, on the estimated river discharge. Based on the spatial-
219 temporal behavior of the reflectance in the analyzed area, a modification of the original formulation is
220 proposed, to take into account the presence of patterns and features resulting from the high spatial
221 resolution images. The new formulation is implemented in the GEE platform for its extensive use.

222 **3.1 Adaptation of the original CM approach to Sentinel-2 data**

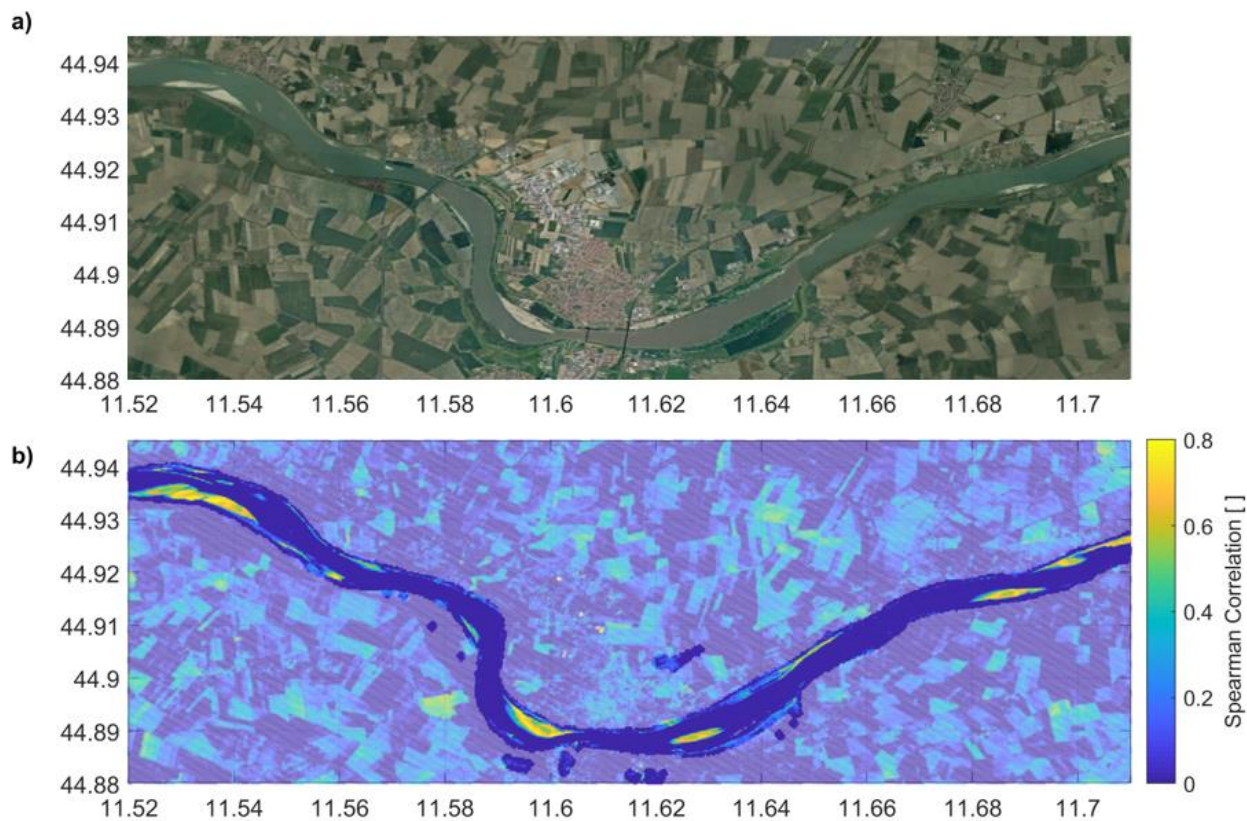
223 Following the previous study of Brakenridge et al. (2007) carried out with AMSR-E datasets, Tarpanelli et al.
224 (2013) developed an approach to estimate river discharge from reflectance data based on the different
225 responses of water and soil to radiation in the NIR band. Indeed, in this region of the electromagnetic
226 spectrum, the reflectance of soil and vegetation is much greater than that of pure water. Therefore, the
227 reflectance of a “wet” pixel, located near the boundaries of the river, changes according to the river
228 discharge variations: specifically, it decreases when river discharge increases and the pixel is flooded.
229 Changes in reflectances are not caused only by changes in river discharge, but could be also related to other
230 sources, such as vegetation, atmosphere composition, soil moisture variations and bidirectional reflectance
231 effects, not negligible over water bodies (Kremezi and Karathanassi, 2019). A “dry” pixel not affected by
232 river discharge variation is then selected to correct the measure, assuming to be affected by the same
233 sources of noise, as mentioned above. By defining the “dry” pixel as Calibration (C) pixel and the “wet” pixel
234 as Measurement (M) pixel, it is possible to obtain a proxy of river discharge by calculating the NIR
235 reflectance C/M ratio. An exponential filter (Wagner et al., 1999) is then applied to the ratio to reduce the
236 effect of residual noises. The locations of C and M are generally calibrated by calculating all the possible
237 combinations of the reflectance C/M ratio on a certain area in the vicinity of the gauged station and by
238 comparing them with the in-situ measurements of river discharge, paying attention that no relevant source
239 of water intake or outlet is located inside the selected region or between the latter and the gauged station.
240 The reflectance ratio providing the highest correlation defines the best location for M and C (see Tarpanelli
241 et al., 2013 for details).

242 The passage from medium to high spatial resolution images required several adjustments. The first one was
243 related to the sensitivity to cloud of the NIR band. The solution adopted by Tarpanelli et al. (2013) for the
244 medium resolution product consisted in discarding all the images affected by clouds. To make the
245 procedure automatic, in this work, the Sentinel-2 images where the label “percentage of cloudy pixels” was
246 greater than 70% were discarded because the residual valid pixels were not statistically sufficient to

247 represent a robust sample. Moreover, the Sentinel-2 cloud probability product was used to mask each
248 Sentinel-2 image's pixels with a probability greater than 50% of being cloudy. The choice of the threshold
249 is based on the fact that for lower values (30-40%), the cloud detection algorithm may occasionally confuse
250 water pixels as clouds in the selected areas, while some cloudy pixels could be confused as valid if higher
251 thresholds were adopted (60-70%). Moreover, just for development purposes, a visual inspection of each
252 NIR images provided by MEEO was carried out to verify the chosen threshold and to discard those in which
253 the river was not clearly visible, i.e. in case of failure of the cloud detection algorithm or in case of few pixels
254 to be used in the analysis.

255 The cloud masking has the side effect to replace clouds with "holes" of no-data in each image. Therefore,
256 the selection of a single pixel for C , as in the original formulation, should be avoided in order to prevent the
257 loss of data in case a cloud appears over the selected "dry" pixel. Following Tarpanelli et al. (2017), the C
258 timeseries was defined as the average reflectance value of the pixels that had a low temporal coefficient of
259 variation (specifically lower than 5th percentile) with respect to all the pixels of the selected area.

260 Regarding the M pixel, as previously noticed, it should be selected over the river boundaries, since this area
261 is the interface between water and soil and it is particularly sensitive to river discharge variations. The
262 selection was constrained to an extended water mask that includes the river and a buffer around it. The
263 European Commission's Joint Research Centre (JRC) Global Surface Water map (Pekel et al., 2016)
264 "max_extent" was used to identify the river region, after the application of a 2-pixels buffer (60 m) applied
265 to enlarge the resulting river area. To assess the best location for M , each pixel within this extended water
266 mask was selected to calculate the C/M ratio (except the pixels already selected for the calculation of C , if
267 any was included in the water mask) over the full available timeseries. Then, the exponential filter was
268 applied, using a value of the filter T parameter equal to 5, which is the median number of days between
269 two consecutive acquisitions of Sentinel-2 images. The Spearman correlation between the observed river
270 discharge and the obtained filtered C/M series was then calculated, as shown in Figure 1b for
271 Pontelagoscuro station. By comparing Figure 1a with Figure 1b, it is evident that the highest correlations
272 for the M pixel coincide with the sediment deposit not vegetated near the river bank.



273

274 *Figure 1: Po River at Pontelagosкуро station. Panel a) represents the location of the area from Google Earth (Copyright ©2021, CNES*
 275 */Airbus, European Space Imaging, Landsat / Copernicus, Maxar Technologies). Panel b) shows Spearman correlation between filtered*
 276 *C/M and the ground observed river discharge, obtained varying the location of pixel M and fixing C as the average of the pixels with*
 277 *coefficient of variation lower than 5th percentile calculated between January 2018 - September 2019.*

278 **3.2 Limits of the CM approach for high resolution products**

279 The expected benefits of the use of satellite imagery at high spatial resolution consist in the possibility to
 280 observe and account for aspects related to river discharge estimation that are not clearly observable from
 281 lower resolution products. In this paper, two main factors were considered to improve the CM approach:
 282 the presence of sediments in the water and the selection of multiple pixels for M . In order to better observe
 283 these phenomena, an analysis was carried out in a smaller area of the Po River at Pontelagosкуро (Figure
 284 2a), within the coordinates [11.588 °E, 11.605 °E, 44.887 °N, 44.9 °N], in the period January 2018-September
 285 2019.

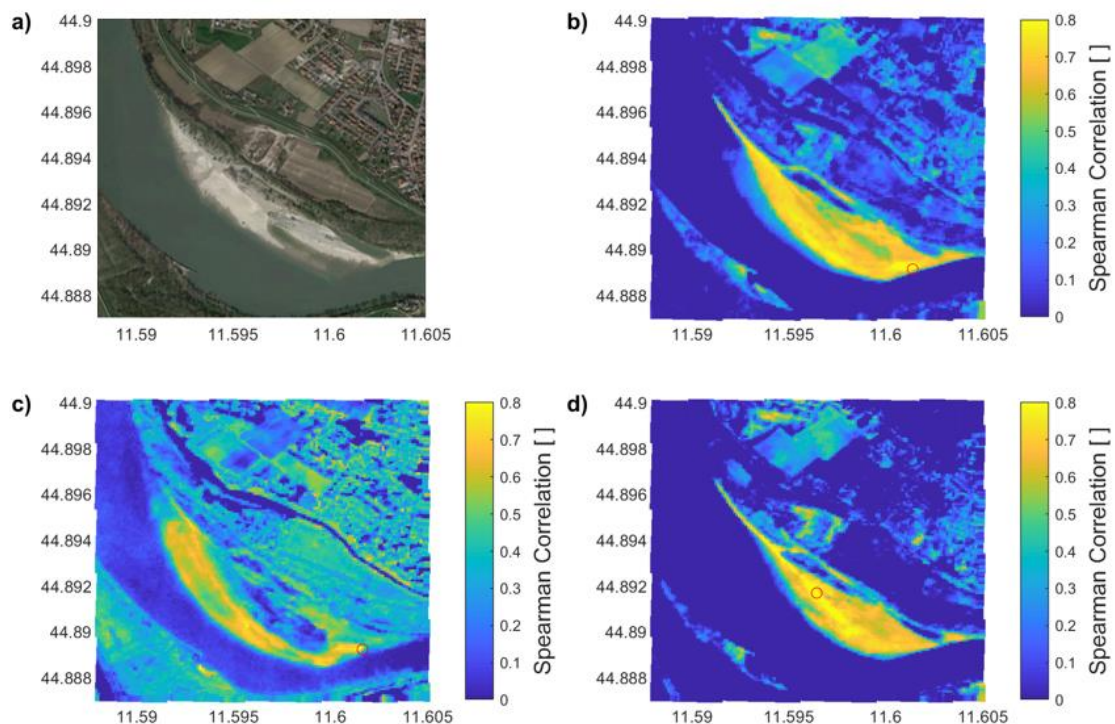
286 The reflectance of clear water in the NIR band is very close to zero, but the reflectance of the river generally
 287 appears to be different from zero, although it remains low compared to soil or vegetation. During high
 288 flows, rainfall events mobilize large amounts of fine sediment into river systems (Keesstra et al., 2019),

289 which contribute to the suspended sediment loads in the river, thus increasing the water reflectance (Ahn
290 and Park., 2020). In these cases, the expected decrease in M reflectance due to inundation is hindered by
291 the reflectance increase caused by the change in suspended sediments in the river. Consequently, the C/M
292 ratio fails in estimating high flow, when the measure is most useful. This phenomenon is clearly visible in
293 Figure 3a, where the hydrograph of the river discharge at Pontelagoscuro during the flood event of October-
294 November 2018 is shown along with the C/M ratio calculated by selecting M as the single pixel that showed
295 the highest Spearman correlation with the observed discharge (see the location of M in Figure 2b). By
296 observation of two images acquired before and in concomitance with the flood event (Figure 3b-3c), the
297 river becomes larger and strongly turbid, increasing its reflectance up to values very close to the
298 surrounding soils (see Figure 3c). During and after the flood event, the suspended sediments decrease due
299 to the catchment sediment dynamic (Keesstra et al., 2019; Gentile et al., 2010), which is different from the
300 river discharge dynamic. Consequently, M decreases and the C/M ratio increases (see Figure 3a), leading
301 to the late detecting of the inundation. In order to correct the issue, particularly important in the
302 hydrograph rising limb, a correction should be introduced in the formulation of the CM approach to taking
303 in account the role of the suspended sediments.

304 Another important aspect is strongly related to the pixel dimension of Sentinel-2. Due to the high frequency
305 of the NIR electromagnetic waves and to the water absorption properties in this band, the NIR reflectance
306 decreases significantly with water depth down to 0.75 m (shallow water condition, Gilvear et al., 2007) and
307 remains constant for lower depths. A completely flooded pixel would therefore lose most of its sensitivity
308 to the water level increasing and thus, to the river discharge variation. A single pixel of Sentinel-2 is too
309 small to account for the variation of the river discharge over the study area because the width of the Po
310 River during a flood event is much larger than the dimension of a single Sentinel-2 pixel (compare Figure 3c
311 and 3b for the Po River width variation). Indeed, by looking at Figure 3a, the C/M ratio obtained from the
312 maximum correlated pixel shows a limited variability during the low flow from July to October 2018, being
313 the chosen M pixel located over completely dry soil. In the high-resolution product analysis, the single pixel
314 can be affected by both long period of dry condition (with reflectance very similar to the soil) alternated

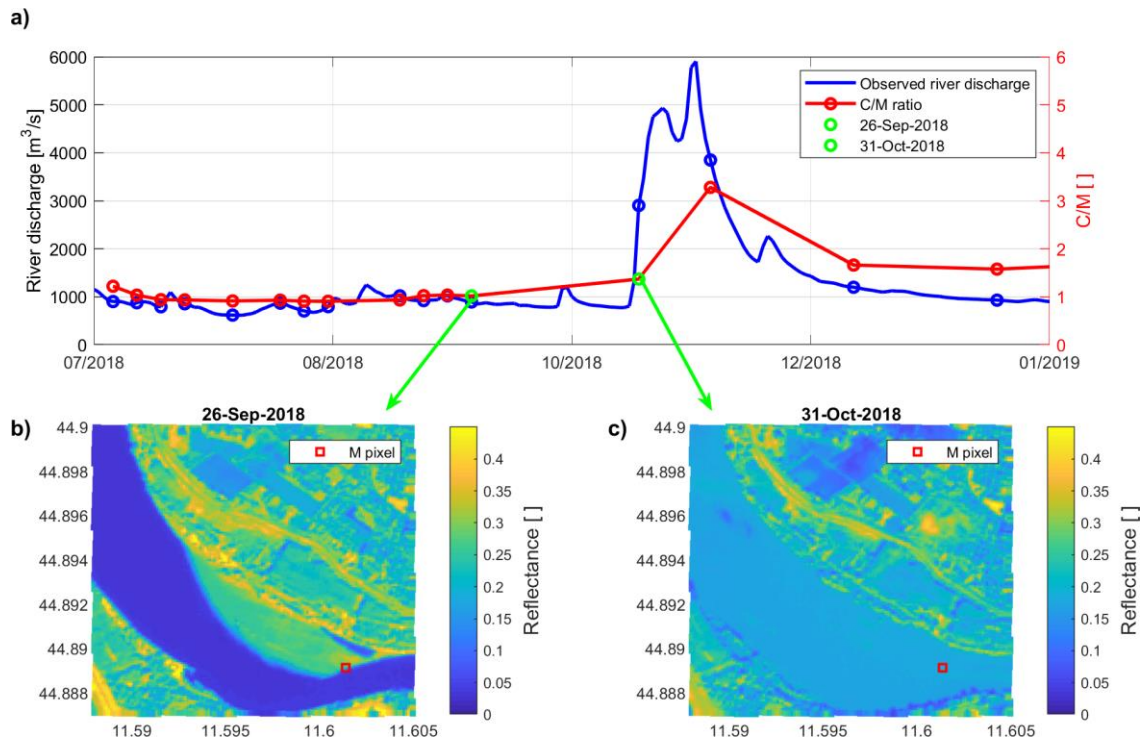
315 with period of saturated conditions (completely flooded pixel, with reflectance very similar to the water
 316 inside the river). When the flow variation causes a pixel to be saturated, the relative C/M ratio stop varying
 317 and therefore the selected pixel cannot be used to properly represent further variation of river discharge.
 318 This is confirmed by the location of the pixels showing the highest correlation coefficients with the ground
 319 observed river discharge: for medium-low flows (river discharge values below the threshold of 50th
 320 percentile, around 1000 m³/s in the selected period), the highest Spearman correlations were obtained
 321 more internally into the river, as shown in Figure 2c. Conversely, the highest Spearman correlations were
 322 obtained in the external area (Figure 2d) for high flow condition (river discharge values above the threshold
 323 of 50th percentile). On the basis of these considerations, the choice of a single pixel for all river conditions
 324 is not advisable: M should be based on multiple pixels, chosen so as to take almost all the river regimes into
 325 account.

326



327

328 *Figure 2: Po River at Pontelagoscuro station. Panel a) represents the area from Google Earth (Copyright ©2021 Immagini ©2021,*
 329 *CNES / Airbus, Maxar Technologies). Panels b), c) and d) show Spearman correlation between filtered C/M and the ground observed*
 330 *total discharge (b), low discharge below the 50th percentile (c), high discharge above the 50th percentile (d). The red circles in panels*
 331 *b, c, d represent the position of the M pixel where the Spearman correlation with the ground-based observations is highest.*



332

333 *Figure 3: Po River at Pontelagoscuoro station. Panel a) shows the hydrograph of the river discharge and the C/M ratio calculated for*
 334 *the analyzed area. The circles on the temporal series represent the days when the satellite overpasses. Panel b) and c) show the*
 335 *Sentinel-2 reflectance images before and during the October-November 2018 flood event, respectively. The red squares in the panels*
 336 *b and c represent the position of the M pixel, where the Spearman correlation with the ground-based observation is the highest.*

337 **3.3 A new formulation: the CMW approach**

338 Following the considerations described in the previous paragraph, two main modifications of the CM
 339 approach are required: the management of suspended sediments effect and the use of multiple pixels for
 340 M selection.

341 The role of suspended sediments is important for a correct evaluation of the M reflectance. As specified
 342 above, the presence of suspended sediments makes the water turbid and increases the M reflectance, with
 343 the consequent decreasing of the C/M ratio especially during high flows. In order to evaluate the
 344 reflectance of the turbid water, it is necessary to average the reflectance of multiple water pixels, W,
 345 located over a region of the river constantly wet (inner part of the river, with water always present, also
 346 during low flows). Assuming W reflectance as a proxy of the presence of suspended sediments, M should
 347 vary between C for low flows (dry conditions) and W for high flows (wet conditions).

348 The original CM approach for the estimation of the river discharge, Q, assumed as:

349
$$Q \propto CM = \frac{c}{M} \quad (1)$$

350 can be then reformulated as:

351
$$Q \propto CMW = \frac{c}{M-W+z} \quad (2)$$

352 where z is a numerical coefficient: in condition of high flow, M tends to be completely flooded and
353 therefore its value is very close to the one of W . Any noise within their measurements can cause their
354 difference being lower than 0, invalidating the obtained index. Furthermore, in case the selected M pixel
355 become completely flooded, the condition of $M = W$ would be verified and the CMW ratio would tend to
356 infinity. For both these reasons, we introduced z as

357
$$z = \max(W - M) + \min(M) \quad (3)$$

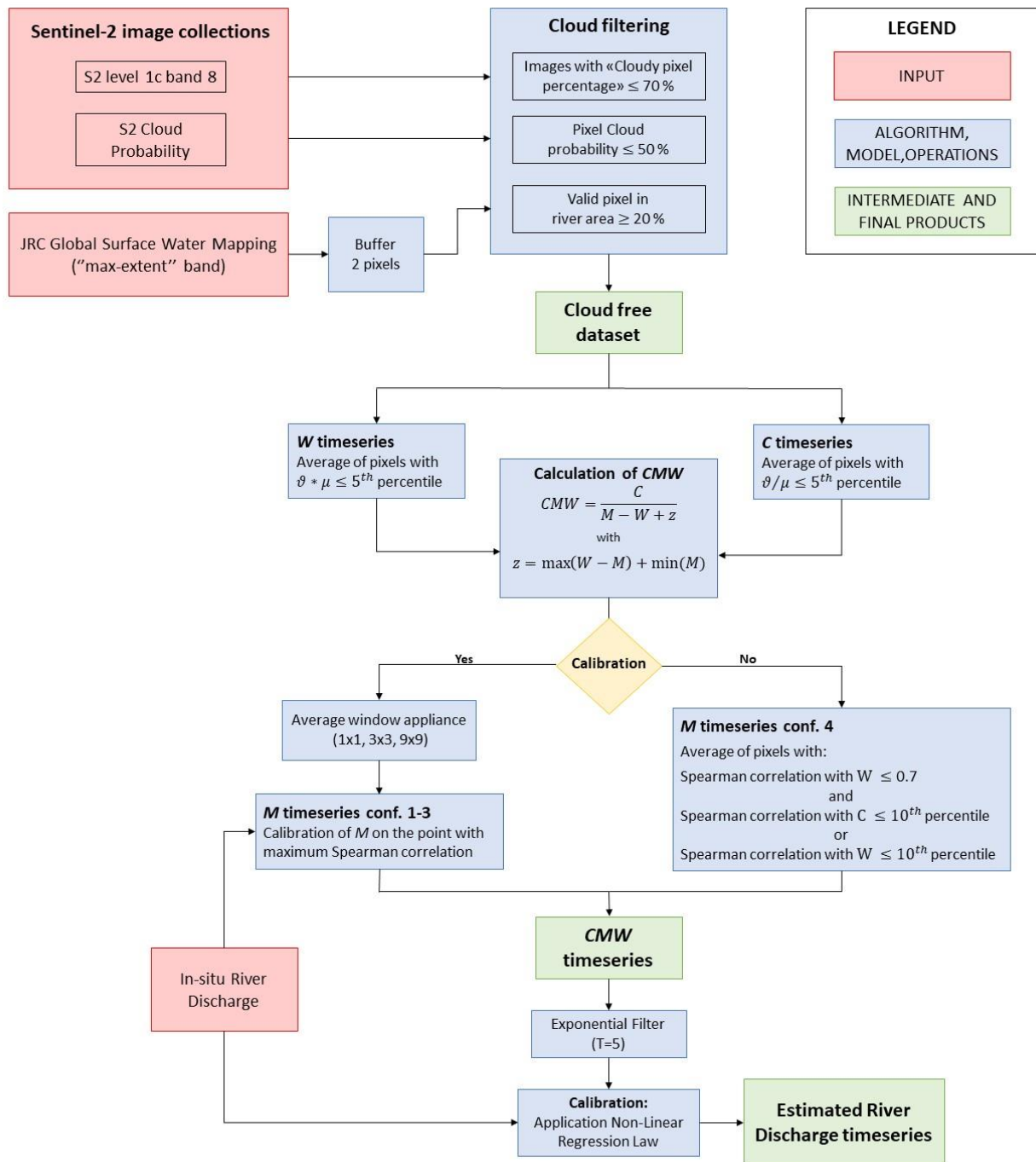
358 W pixels have to be selected within the river. This condition can be ensured by considering only the pixels
359 included within a water mask, and by calculating, for each pixel, the product between the temporal average
360 of the reflectance (the lower, the more probable presence of water) and its temporal standard deviation
361 (the lower, the less variation occurs in the pixel). The W reflectance of eq. (2) can be then calculated by
362 averaging the NIR reflectance of the pixels for which the product above defined is below a threshold (fixed
363 at the 5th percentile according the sensitivity analysis, see Section 4.1), in order to identify the pixels where
364 water is always present.

365 The timeseries of M should be obtained by averaging reflectances from multiple pixels located in an area
366 within the water mask that is particularly sensitive to variations in river discharge. In this paper, we
367 investigated several configurations to analyze the best strategy: 1) single pixel; 2) 3x3 pixels; 3) 9x9 pixels;
368 4) multiple pixels. In particular, configuration 1) was investigated to be compared to the original
369 formulation. Configurations 2) and 3) were investigated to identify possible improvements from the
370 evaluation of multiple pixels. Finally, configuration 4) was included in order to not calibrate the position of
371 the M pixels with the ground observed discharge timeseries, but to approach an automated procedure for
372 selecting an area or set of pixels that is particularly suitable for describing the river discharge variation. This

373 area was identified following the fact that the NIR reflectance timeseries of the constituent pixels should
374 not be correlated with the W and/or C timeseries. Specifically, pixels sensitive to low flow should be poorly
375 correlated with C (because they are inundated most of the time), while those sensitive to high flow should
376 be poorly correlated with W (because they are dry most of the time). Therefore, the condition that a pixel
377 has to meet for being selected as one of the optimal M pixels is a very low Spearman correlation in
378 reflectance with the pixels of W or C (lower than the 10th percentile, according to the sensitivity analysis
379 described in Section 4.1). In addition, to avoid selecting permanently flooded pixels (because they are
380 poorly correlated with C), a maximum threshold on the pixels Spearman correlation with W is set to a fixed
381 value (0.7 according the sensitivity analysis, see Section 4.1).

382 Summing up, the new procedure to estimate river discharge from high-resolution NIR reflectance data
383 consists into the following steps (see also Figure 4):

- 384 1. The water mask is delineated through the “max extent” from JRC Global Surface Water map by considering
385 a buffer of 3 pixels around each pixel of the water map;
- 386 2. each image where the field “percentage of cloudy pixels” of the Sentinel-2 product indicates a percentage
387 of cloudy pixels greater than 70% is directly discarded and each image’s pixels in which the cloud probability
388 is greater than 50% is masked out. Furthermore, the images with less than the 20% (empiric threshold) of
389 valid pixels over the extended water mask are also discarded;
- 390 3. the C timeseries is calculated by averaging the reflectance of those pixels showing a temporal coefficient
391 of variation lower than a certain threshold (5th percentile, see Section 4.1);
- 392 4. the W timeseries is calculated by averaging the reflectance of those pixels, included in the water mask (step
393 3), whose product between temporal average and temporal standard deviation of reflectance values is
394 below a certain threshold (5th percentile, see Section 4.1);



395

396 *Figure 4: Flowchart of the CMW approach. The thresholds are identified according to the analysis shown in Section 4.1.*

397 5. the M timeseries are obtained by considering the four configurations mentioned above. In the first three
 398 configurations, each pixel within the water mask (step 3) is used to calculate the ratio CMW according to
 399 eq. (2): in the case of configuration 2 and 3 each pixel value is resampled with a moving average spatial
 400 window of 3x3 pixels and 9x9 pixels, respectively. The pixel where the Spearman correlation between CMW

401 and the ground observed discharge is maximized, is selected as M pixel. Alternatively, for configuration 4,
402 multiple pixels in the water mask having Spearman correlation with W or C below a percentile threshold
403 (10th percentile, see Section 4.1) and with W below a fixed value (0.7, see Section 4.1) are selected and
404 averaged to obtain M without the use of any information from ground data (for this reason we referred to
405 this configuration as uncalibrated).

406 6. C , W and M are used to calculate CMW index, which is then filtered through the exponential smoothing
407 function with T equal to 5 days.

408 In order to compare the improvements of the new CMW approach, the CM approach was also tested,
409 following the same steps except for the 4th, because W was not needed, and replacing CMW index by C/M
410 ratio in the step 5.

411 Both approaches were also tested using MODIS data, in order to compare the high with the medium spatial
412 resolution data and evaluate potential benefits. The adopted procedure is very similar to the Sentinel-2
413 one, with the following exceptions:

414 1) a buffer of 3 pixels (instead of 2) is applied to the Global Surface Water map (Pekel et al., 2016) to
415 identify the extended area over the Tiber River, since otherwise the water mask size could not reach
416 the spatial resolution of MODIS pixels;

417 2) due to the absence of a MODIS Cloud Probability product in GEE, the cloud filtering is performed by
418 exploiting the “state_1km” band of MYD09GA product. Specifically, all the pixels labelled as cloudy,
419 mixed, cirrus, snow, ice or adjacent to cloud are masked out;

420 3) configurations 2 and 3 are not applied to MODIS data because a single pixel of this sensor exceed the
421 sensitivity of a Sentinel-2 pixel in configuration 3 (~90x90 m of Sentinel-2 against ~250x250 m of
422 MODIS).

423 **3.4 Performance scores**

424 In terms of performance scores, the coefficients of correlation of Spearman, R_s , and Pearson, R_p , were
425 calculated between the CMW or CM timeseries and the ground observations of river discharge. Then, the

426 estimation of the river discharge from *CMW* or *CM* was performed by a fitting function (cubic function)
427 described, for *CMW*, as:

$$428 \quad \tilde{Q} = a + b \cdot CMW + c \cdot (CMW + d)^3 \quad (4)$$

429 where \tilde{Q} is the estimated river discharge, a , b , c and d are parameters, with b and c being positive so as to
430 ensure the monotonicity of the function. The parameters were calibrated for each station using in-situ
431 discharge observations in order to obtain the most accurate river discharge estimation.

432 Once the function was defined, three performance scores were computed to quantify the error in the final
433 estimation of the river discharge: i) the Root Mean Square Error (*RMSE*), ii) the relative *RMSE* (*rRMSE*),
434 defined as the ratio between *RMSE* and the mean observed river discharge and iii) the Nash-Sutcliffe
435 efficiency (Nash and Sutcliffe, 1970), *NS*.

436 **4. Results and discussions**

437 This section contains the results obtained from the analyses of Sentinel-2 and MODIS data, carried out with
438 the help of the Google Earth Engine, in the period: July 2015 – December 2020. First, the sensitivity analysis
439 carried out on the thresholds selected to identify the masks for every configuration is shown for
440 Pontelagoscuro station, along the Po River. Then, taking in account the selected thresholds, the approach
441 is applied at Pontelagoscuro and Montemolino (Tiber River) stations and a deep analysis on the masks and
442 the consequent results is carried out. Finally, the analysis is extended to the seven stations and the results
443 are shown by comparing the original algorithm *CM* and the new formulation *CMW* for both the products
444 Sentinel-2 and MODIS.

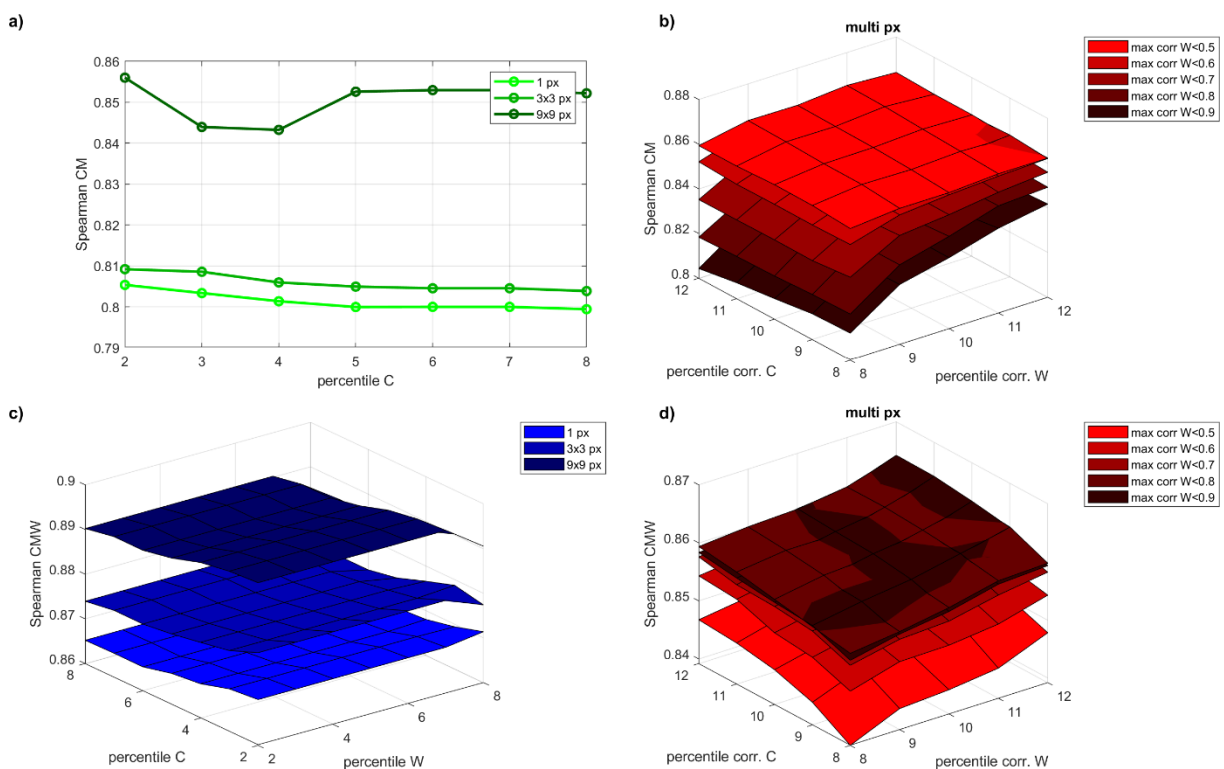
445 **4.1 Sensitivity analysis**

446 The methodology developed to infer the river discharge from NIR reflectance is based on the selection of
447 multiple thresholds, needed to obtain the static masks of C , W and M . In order to assess the validity of the
448 selected thresholds, a sensitivity analysis was conducted at the station of Pontelagoscuro. Regarding the

449 calibrated configurations (1-3), the percentile thresholds to select *C* and *W* pixels was made to vary
450 between the 2nd and the 8th percentile (1 percentile step), resulting in 21 possible layouts for the *CM*
451 approach (7 *C*-masks * 3 calibrated configurations) and 147 possible layouts for the *CMW* approach (7 *C*-
452 masks * 7 *W*-masks * 3 calibrated configurations). Regarding configuration 4, the thresholds related to the
453 *C* and *W* masks selection were fixed at the 5th percentile (based on the results of the other configurations)
454 and just the ones related to the *M* selection were modified, to limit the degrees of freedom of the analysis.
455 Therefore, the percentile correlation with *C* and *W* was made varied between the 8th and the 12th percentile
456 (1 percentile step) and the maximum correlation with *W* between 0.5 and 0.9 with a (0.1 step), resulting in
457 125 different layouts (5 *C* percentile limits * 5 *W* percentile limits * 5 maximum correlation with *W*
458 thresholds). The sensitivity of the thresholds was evaluated in terms of Spearman correlation *R*s calculated
459 between the obtained *CM* or *CMW* timeseries and the observed river discharge.

460 The sensitivity analysis was carried out for the station of Pontelagoscuro, where the methodology was
461 developed. The obtained results in terms of *R*s are shown in Figure 5. Regarding the calibrated approaches
462 (configurations 1 to 3, Figure 5a and 5c), the selection of the percentiles for obtaining the *C* and *W* masks
463 has little incidence on the performance of *CM* and *CMW*, varying the *R*s in the range of 0.01. Based on these
464 evidences, the percentile thresholds to select *C* and *W* pixels was selected equal to the 5th percentile,
465 according to the work of Tarpanelli et al., 2017. The variability of configuration 4 results (Figures 5b and 5d)
466 is instead in the order of 0.06 for *CM* approach and 0.03 for *CMW* approach, mostly ascribable to the
467 threshold related to the maximum correlation with *W*. When the simple *CM* approach is considered (panel
468 5b) the high *W* correlation thresholds (dark red surfaces) allow to select many pixels in the inner river well
469 correlated with *W*, providing worse performances with respect to the low thresholds (light red surfaces)
470 due to the fact that these pixels are flooded most of the time and thus particularly sensitive to the sediment
471 load of the river. This effect is worsened by choosing a low percentile correlation with *W* or an high
472 percentile correlation with *C*: in both cases, the portions of pixels sensitive to low flows within *M* tends to
473 be greater than those sensitive to high flows, obtaining the same aforementioned effect due to the absence
474 of the sediment contribution within the *CM*. Conversely, if the *CMW* approach is applied (panel 5d), the

475 presence of sediments is taken in account and accepting a larger number of pixels well correlated with W
 476 becomes advantageous for the final results. In the Pontelagoscuero area, positive effects on the results can
 477 be obtained also by increasing the percentile correlation thresholds of C and W, due to the presence of the
 478 vast area subjected to periodic floods noted in Figure 2. Notwithstanding this, such an area is not
 479 encountered in all the sections of a river nor in all the rivers. Therefore, the M mask selection thresholds
 480 for the configuration 4 were fixed at the 10th percentile for the correlation with W and C, whereas the
 481 maximum correlation with W was set at 0.7.



482

483 *Figure 5: Sensitivity analysis at Pontelagoscuero station (Po River) in terms of Rs between CM and CMW timeseries and ground based*
 484 *river discharge. The subplots on the left show the Rs obtained using the CM (panel a) and CMW (panel c) approaches by varying the*
 485 *C mask from the 2th to the 8th percentile (panel a) and the C and W mask from 2th to 8th percentile (panel c) for the configurations 1*
 486 *to 3 (indicated from light to dark green or blue). The subplots on the right show the Rs obtained by using the CM (panel b) and CMW*
 487 *(panel d) approaches for the uncalibrated configuration 4 where the correlation with C and W is made to vary between the 8th and*
 488 *the 12th percentile and the maximum correlation with W between 0.5 and 0.9.*

489 **4.2 Pontelagoscuero results assessment**

490 Focusing on Pontelagoscuero station, a total of 743 images of Sentinel-2 Level 1C band 8 were available on
 491 GEE (accessed on 03/03/2022) during the period July 2015 – December 2020, obtained in 316 different

492 days. After the cloud masking procedure, 168 cloud free images were retained and analyzed through the
493 *CM* and the *CMW* procedure.

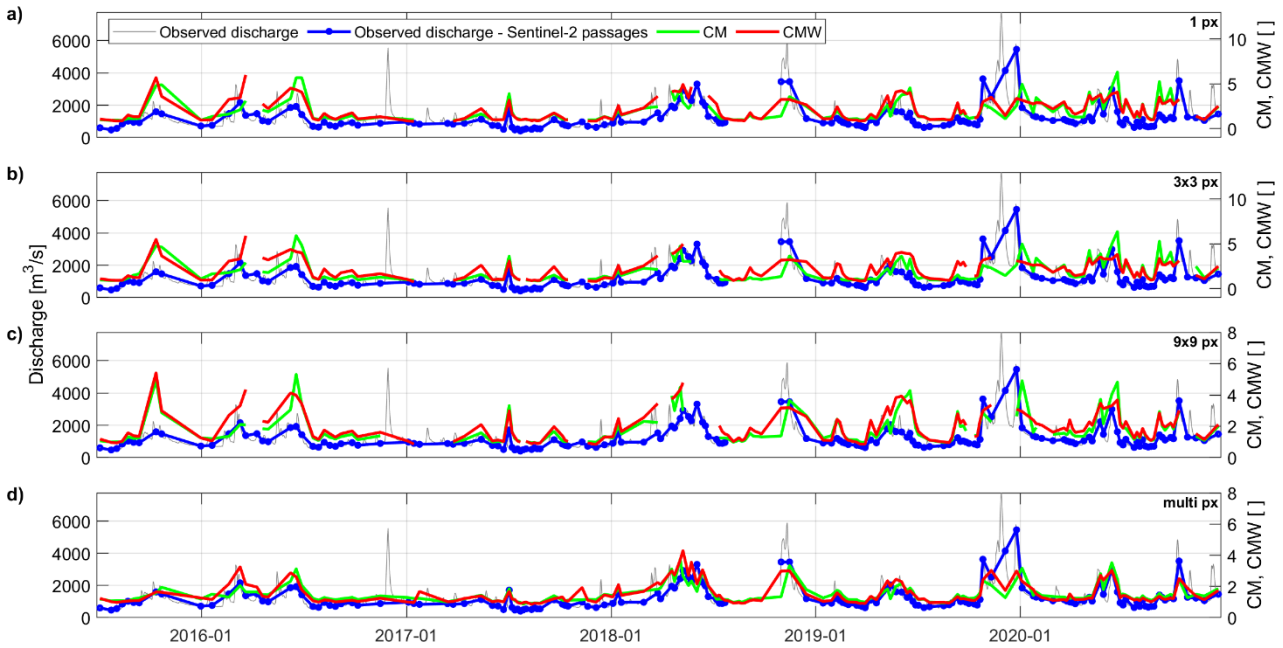
494 The timeseries obtained for the four configurations are shown in Figure 6, together with the hydrograph of
495 the observed river discharge. All the configurations were successful in detecting the river dynamic. *CMW*
496 approach showed a general better agreement with the observed discharge with compared to the *CM*
497 approach, in particular at the beginning of the flood events in March 2016, May 2018, November 2018 and
498 November 2019, when the high river discharge triggered a high level of sediment transportation. This can
499 be noticed by observing the scatter plots on Figure 7, representing the *CM* and the *CMW* timeseries against
500 the observed river discharge in green and red, respectively: most of the green dots in the upper-left area
501 of the panels, characterized by low reflectance ratios and high discharge values due to the high sediment
502 transportation condition, are avoided with *CMW* approach.

503 The results of the different configurations show that by increasing the number of pixels for the calculation
504 of *M* (from the first to the fourth configuration, respectively on the top and the bottom of the Figure 6),
505 *CM* approach improves, gradually agreeing with *CMW* approach with the exception of the aforementioned
506 flood events. Indeed, by selecting a greater number of pixels, the reflectance ratio is more sensitive to river
507 discharge variation, since each pixel becomes wet for different flow conditions and thus a greater variability
508 of the ratio is obtained. Moreover, selecting a large area makes the approach less sensitive to residual
509 sources of error, e.g., residual clouds and cloud shadows, since it is unlikely that all the *M*-mask selected
510 area is equally affected by the noises.

511 Even though the new formulation is in good agreement with the observed river discharge and it improves
512 the previous methodology, the peak floods experienced in November 2018 and November 2019 were still
513 underestimated for all the configurations. This is due to the fact that the *M* pixels selected in the
514 configurations were completely flooded for values of river discharge lower than those of the two peaks,
515 preventing them to give information of flow events much greater than those at which they were saturated.
516 Moreover, a registration issue between Sentinel-2 images was detected, leading to the impossibility of

517 selecting precisely the river boundary pixels (additional details in Section 4.3). Indeed, the expected river
518 width variation for very high flow is limited and the pixels sensitive to it are extremely few, due to the river
519 embankments presence. The inaccuracy of Sentinel-2 imagery registration and the low frequency of highest
520 discharge peaks (encountered just two times during the study period) meant that none of the pixels
521 affected by them were selected by the M configurations. These flood events had in fact too low importance
522 in terms of Spearman correlation to influence the pixels selection of the calibrated approaches and too low
523 statistical significance to affect the one of the uncalibrated one.

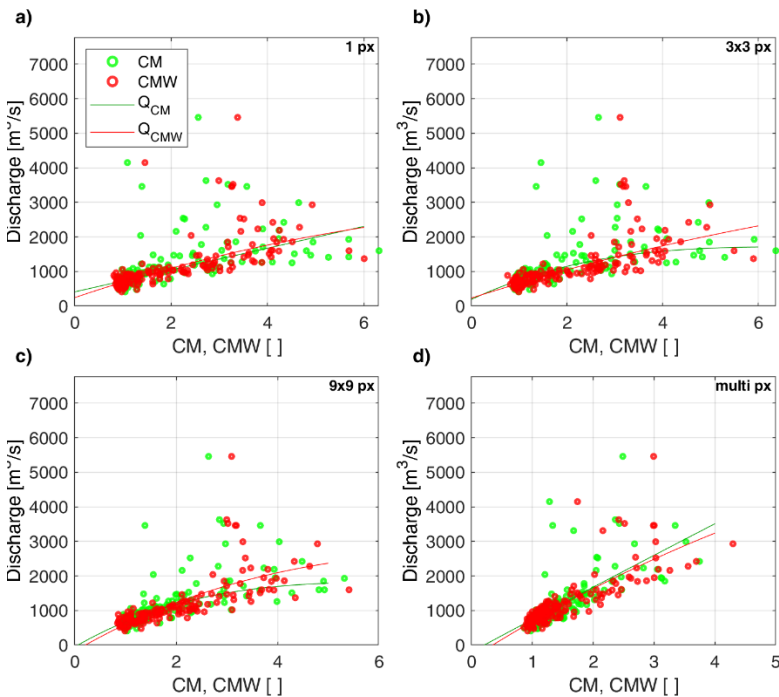
524 It should be finally noticed that additional river discharge peaks were not detected due to the low temporal
525 resolution of Sentinel-2 (no overpass during the flood events) and the discard of the images affected by
526 clouds (gray peaks in Figure 6, e.g., see the flood event in December 2016). Notwithstanding this, the
527 simulated discharges were in good agreement with the ground observed discharges, as confirmed by the
528 performance indices of Table 1, that shows correlations R_p higher than 0.62 and R_s up to 0.89. If compared
529 with the CM approach, R_p was higher than the CMW approach for all the configurations, with increments
530 from +0.09 to +0.14. Similarly, improvements were observed by the use of multiple pixels rather than a
531 single one, from +0.02 to +0.22 by considering both approaches together, in terms of Pearson correlation.
532 Slighter improvements were also observed in terms of R_s . In terms of river discharge, by applying the cubic
533 law of eq. 4, the performance scores of Table 1 show a decrease of $rRMSE$ up to 0.19 and an increase of NS
534 up to 0.45, considering both the multiple pixels and the different methodology.



535

536 *Figure 6: Po River at the Pontelagoscuo station: comparison in terms of temporal series between the ground observed discharge and*
 537 *the CM (green) or CMW (red) timeseries for configuration 1 (single pixel, panel a), configuration 2 (3x3 averaged pixel, panel b),*
 538 *configuration 3 (9x9 averaged pixel, panel c) and for configuration 4 (multiple uncalibrated pixels, panel d). The observed discharge*
 539 *is shown in light grey for the full timeseries and in blue for the measurements coincident to Sentinel-2 overpass.*

540



541

542 *Figure 7: Po River at the Pontelagoscuo station: comparison in terms of scatter plots between the ground observed discharge and*
 543 *the CM (green) or CMW (red) timeseries for configuration 1 (single pixel, panel a), configuration 2 (3x3 averaged pixel, panel b),*
 544 *configuration 3 (9x9 averaged pixel, panel c) and for configuration 4 (multiple uncalibrated pixels, panel d). The calibrated*
 545 *relationships of the estimated river discharge are shown as solid lines for both CM (dark green) and CMW (red).*

546 *Table 1: Performance scores for the different configurations (1 for single pixel, 2 for 3x3 pixels, 3 for 9x9 pixels and 4 for multiple*
 547 *pixels) for the period 2015-2020 at Pontelagoscuo station. Spearman (R_s) and Pearson (R_p) correlations are calculated between the*
 548 *CMW or CM timeseries and the ground observed river discharge. Root Mean Square Error (RMSE), relative Root Mean Square Error*

549 (*rRMSE*) and Nash-Sutcliffe efficiency (*NS*) refer to the comparison between the simulated discharge calculated according eq. 4 and
 550 the ground observed discharges.

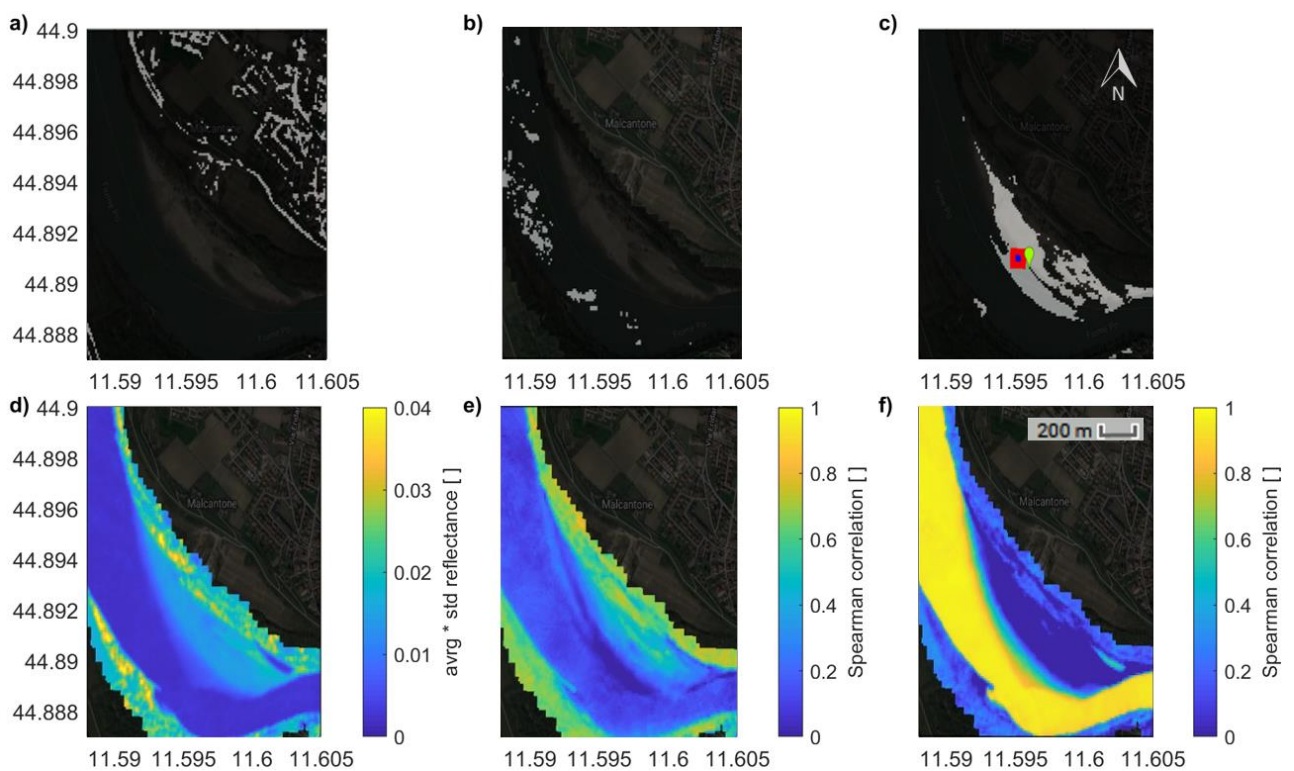
Config.	CM					CMW				
	<i>Rs</i>	<i>Rp</i>	<i>RMSE</i>	<i>rRMSE</i>	<i>NS</i>	<i>Rs</i>	<i>Rp</i>	<i>RMSE</i>	<i>rRMSE</i>	<i>NS</i>
	[-]	[-]	[m ³ /s]	[-]	[-]	[-]	[-]	[m ³ /s]	[-]	[-]
1	0.80	0.48	689	0.57	0.17	0.87	0.62	600	0.50	0.38
2	0.80	0.53	655	0.54	0.26	0.88	0.64	556	0.47	0.40
3	0.85	0.61	581	0.49	0.35	0.89	0.70	508	0.43	0.50
4	0.85	0.70	545	0.45	0.48	0.86	0.79	461	0.38	0.63

551

552 By moving from the single to aggregated pixels, the performance increased, reaching the best values for
 553 configuration 4. This is justified by the Sentinel-2's pixels size, which is smaller than the Po River width: it is
 554 convenient to aggregate multiple pixels for a proper representation of the river discharge.

555 Figure 8a shows the pixels of the box with lower (less than 5th percentile) temporal coefficient of variation,
 556 selected to obtain the *C* timeseries. As expected, they are mostly concentrated over roads and bare soil
 557 areas, due to the temporal stability of the reflectances measured in these areas. Figure 8b shows the pixels
 558 within the water mask selected for the *W* timeseries. They represent the lower (less than 5th percentile)
 559 product between the average reflectance and its standard deviation, represented in Figure 8d, and they are
 560 correctly identified in the center of the river. Figure 8c represents the selected *M* pixels: the different
 561 calibrated configurations (1-3) are highlighted with different colors, while the light grey area shows the *M*
 562 pixels selected by the configuration 4 (without using any ground observation for calibration) and coming
 563 from the evaluation of the Spearman correlation between the reflectance timeseries of each pixel included
 564 in the water mask (except those pixels already used for *W* and *C*) and the timeseries already calculated for
 565 *C* (or *W*) deriving from the masks in Figure 8e (or 8f). Specifically, the blue pixels in Figure 8e are poorly
 566 correlated with *C*, and coincide with those showing greater correlation with low flow (compare with Figure
 567 2c) and those always inside the water. Conversely, the pixels in yellow in Figure 8f are highly correlated
 568 with *W* timeseries (stable water pixels) and therefore identify the inner river, while the blue pixels coincide

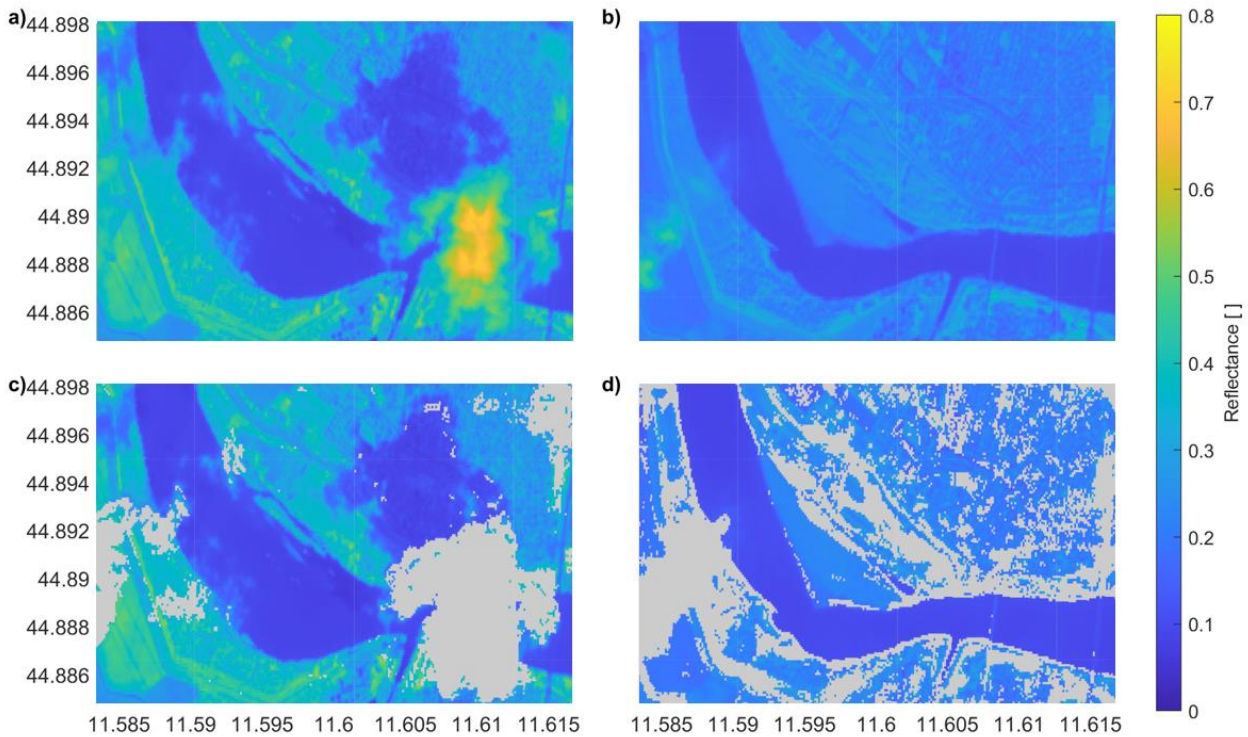
569 with those showing greater correlation the high flow (compare with Figure 2d). Permanently flooded pixels
 570 (highly correlated with W) are also poorly correlated with C (Figure 8e), therefore the condition of
 571 Spearman correlation with $W < 0.7$ was needed to avoid their selection for the M -mask in configuration 4.
 572 This condition helps reaching a tradeoff between keeping low the number of pixels selected for M and
 573 ensure the sensitivity of the procedure to the high river discharge. In fact, the pixels sensitive to high flow
 574 condition are expected to be less in number than those sensitive to low flow condition due to the nature
 575 of the phenomenon: therefore, it is important to limit the selection of the pixels poorly correlated with C ,
 576 sensitive to low flow, in order to balance the M pixels distribution. Despite this step, the M -mask obtained
 577 by configuration 4 is much larger than those obtained by configurations 1-3. However, this allows to
 578 decrease the sensitivity of M to atmospheric noises, by diluting the pixels affected by random clouds with
 579 a larger number of pixels.



580
 581 *Figure 8: Po River at Pontelagoscurio: in light gray the pixels used for calculating C (panel a), W (panel b). Panel c shows the location*
 582 *of the pixel M obtained from configuration 1 (1 pixel, green dot), configuration 2 (3x3 pixels, blue square), configuration 3 (9x9 pixels,*
 583 *red square) and configuration 4 (multiple uncalibrated pixels, light gray). Panel d) shows the product between the average and*
 584 *standard deviation of the NIR reflectance for each pixel within the water mask; panel e) and f) represent the Spearman correlation*
 585 *map between the reflectance timeseries of each pixel in the water mask and the average reflectance calculated for the C (panel e)*
 586 *and W timeseries (panel f), respectively. Background Copyright ©2021 Immagini ©2021, CNES / Airbus, Maxar Technologies.*

587 In order to better assess the source of errors in terms of river discharge, we analyzed the Sentinel-2 images
588 acquired during the days in which the difference between estimated and observed river discharge was the
589 highest. One of the main disagreement reasons is the clouds presence. Despite the Sentinel-2 cloud
590 probability mask allows to remove the majority of the them, some residual clouds were undetected and
591 unmasked. Furthermore, cloud shadows were not considered in the masking process, and the pixels
592 affected by those were not removed. Clouds are very bright in the NIR, therefore their presence cause river
593 discharge overestimation when they overlap with C or W and underestimation when they coincide with M
594 pixels. In contrast, cloud shadows reduce the reflectance of the affected pixels, causing underestimations
595 when they coincide with C or W and overestimations when they overlap with M . An example of these
596 issues is shown in Figure 9. The panels a and c show the Sentinel-2 1c band 8 image collected on 24/05/2018
597 over Pontelagoscuro area, before and after the application of the cloud probability mask at 50%. Residual
598 clouds can be clearly distinguished over the river, while a cloud shadow is clearly distinguishable in the top
599 of the figure. Panel b and d of Figure 9 show the image collected on 10/06/2017: few clouds were present
600 on the west area of the river, but the algorithm masked out a large number of pixels that were confused
601 with clouds, hampering the river discharge estimation. This is another source of noise cloud-related: in
602 some cases, the mask fails in correctly identifying the clouds, filtering out valid pixels.

603 Another issue is related to the reflectance values of C . It was noticed that the reflectance values of the
604 terrain pixels were occasionally significantly lower/higher than in other days, potentially due to, e.g., the
605 effect of soil moisture variations or bidirectional reflectance effects. These conditions led to a lower/higher
606 C value, which caused the under/overestimation of CMW and consequently of the flow peaks, worsening
607 the effect of the pixels' saturation. This is the case for the floods occurred at the ends of 2020 (Figure 6),
608 where CM and CMW index underestimated the river discharge due to a low value of C , causing most of
609 the CMW outliers that can be noticed in Figure 7. Finally, it should be noticed that the fitting functions
610 shown in Figure 7 were supposed to exhibit steeper behavior to compensate the effect of the M pixels
611 saturation. The presence of residual clouds and shadow, however, generated outliers that compromised
612 the effectiveness of the relationship, highlighting the need to improve the masking procedure.



613

614

615

Figure 9: Po River at Pontelagoscuro: Sentinel-2 1c Band 8 images acquired on 24/05/2018 (Panel a and c) and on 06/10/2017 (Panel b-d), prior (a and b) and after (c and d) the application of the cloud mask equal or above 50% of cloud probability (light gray area).

616

4.3 Montemolino results assessment

617

The results obtained at Montemolino station, over the Tiber River, are described in this section. For this

618

area the total number of available images for the period 2015-2020 was 728, obtained in 625 different days.

619

After the cloud masking, 373 images with valid data were analyzed (based on the access on 03/03/2022).

620

Figure 10 shows the pixels selected from each mask at the Montemolino station over the Tiber River. As for

621

Pontelagoscuro station, *C*-mask (Figure 10a) coincided with roads, parking and bare soil. The river bed was

622

again successfully identified by the product of the average reflectance and its standard deviation (Figure

623

10d). The holes within the river course were caused by river zones that were not detected from the JRC

624

Global Surface water map. In Figure 10b, the pixels selected for *W* were shown. It can be noticed that they

625

did not fall within the center of the river but prevalently on its southern borders. A possible cause of this

626

phenomena is the shadow created by the riparian vegetation, which decreases the average reflectance of

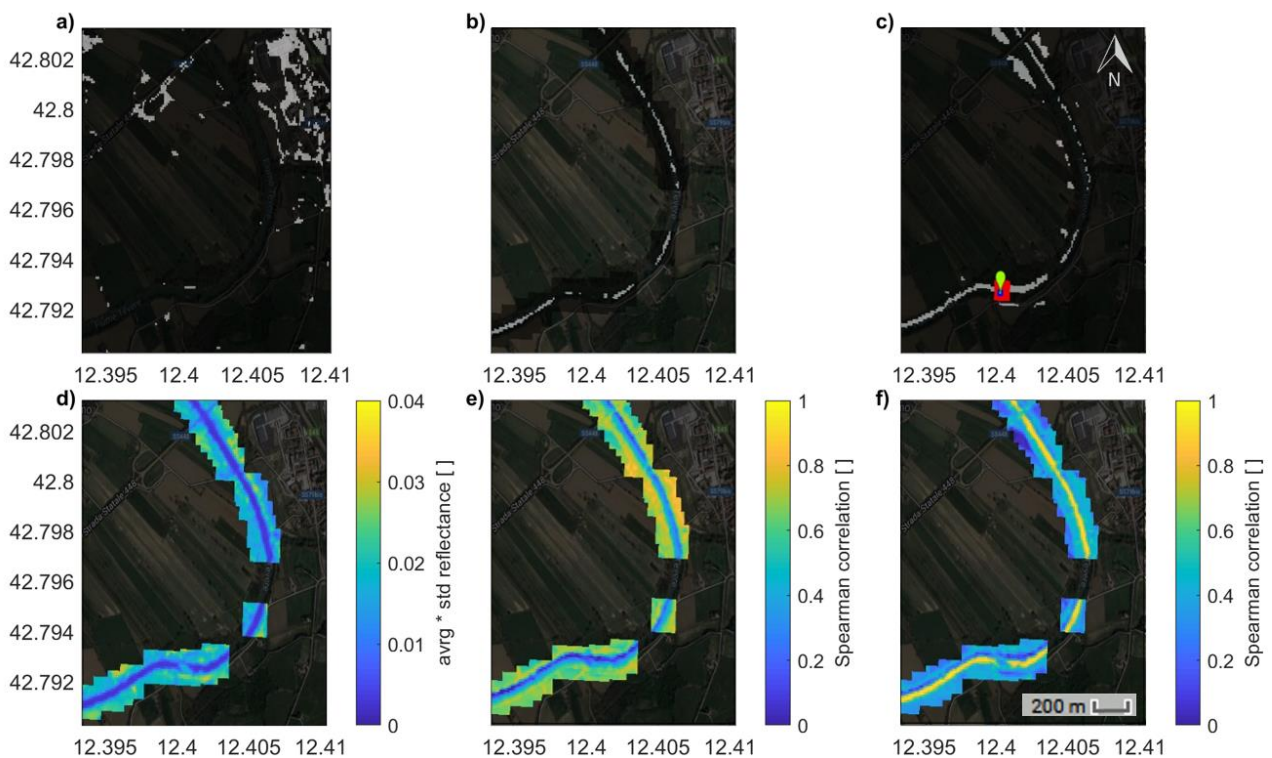
627

the pixels affected by it and facilitates their selection. Figures 10e and 10f show the Spearman correlation

628

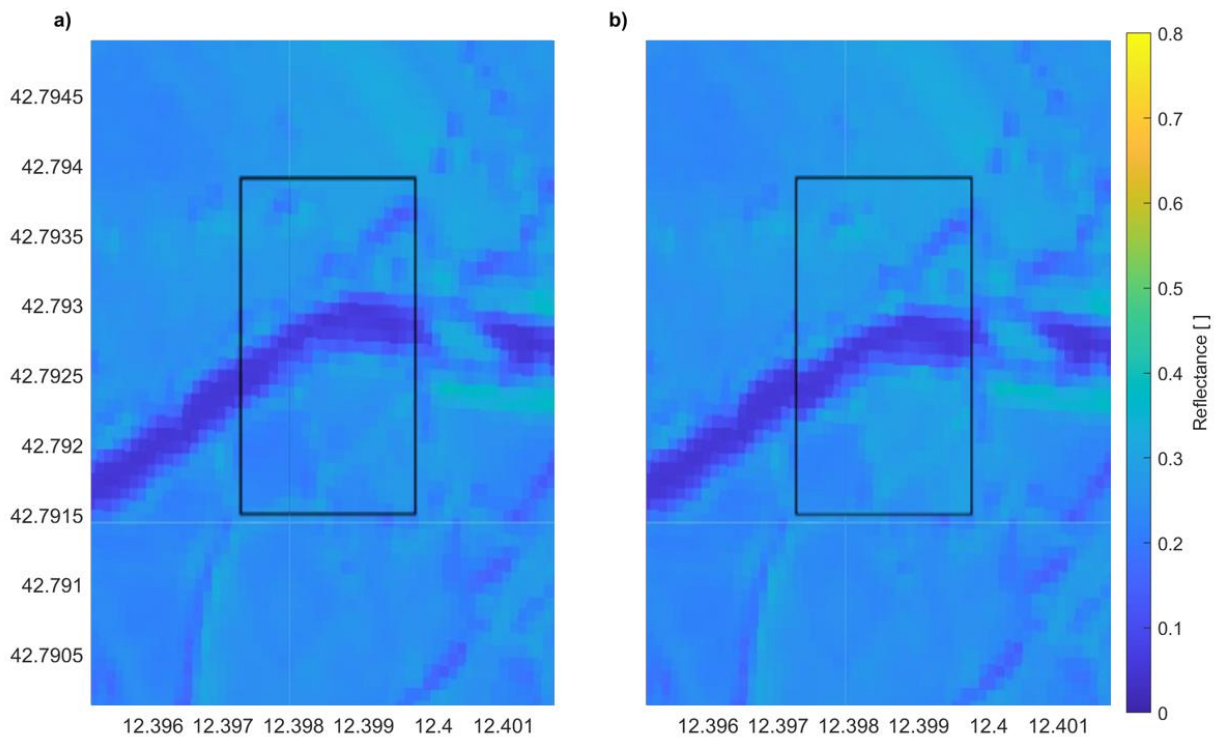
between the reflectance timeseries of each pixel in the water mask and the *C* and *W* timeseries,

629 respectively. Figure 10e shows many pixels poorly correlated with C over the northern river boundaries of of
 630 the southern river branch, which were selected to obtain M timeseries by the uncalibrated configuration,
 631 as shown in Figure 10c. In contrast with the results obtained at Pontelagoscuro, by comparing Figure 10e
 632 with Figure 10f, it can be noticed that many of these pixels show a low correlation with W as well, while for
 633 the Po River the pixels poorly correlated with C were better correlated with W and vice versa. By further
 634 analyzing the selected area, a failure of the image co-registration procedure was detected: the coordinates
 635 of each image often not correspond to the real coordinates of the area, but there was often a
 636 misregistration of a few pixels. An example of this issue is shown in Figure 11, where the Sentinel-2 1c band
 637 8 image acquired on 22/08/2017 is shown in panel a along with a clipping of the image acquired on
 638 27/08/2017 in panel b: a miss-registration of ~ 1 -2 pixels can be noticed by comparing the two images.



639
 640 *Figure 10: Tiber River at Montemolino: in light gray the pixels used for calculating C (panel a), W (panel b). Panel c shows the location*
 641 *of the pixel M obtained from configuration 1 (1 pixel, green dot), configuration 2 (3x3 pixels, blue square), configuration 3 (9x9 pixels,*
 642 *red square) and configuration 4 (multiple uncalibrated pixels, light gray). Panel d) shows the product between the average and*
 643 *standard deviation of the NIR reflectance for each pixel within the water mask; panel e) and f) represent the Spearman correlation*
 644 *map between the reflectance timeseries of each pixel in the water mask and the average reflectance calculated for the C (panel e)*
 645 *and W timeseries (panel f), respectively. Background Copyright ©2021 Immagini ©2021, CNES / Airbus, European Space Imaging,*
 646 *Maxar Technologies.*

647



648

649

650

Figure 5: Tiber River at Montemolino: detail of Sentinel-2 1c Band 8 image acquired on 22/08/2017 (Panel a and b) and on 27/08/2017 (central patch Panel b), showing a miss-registration between the two images of about 1-2 pixels.

651

This apparent movement of the full river section caused many pixels vary “artificially” between soil and

652

water, which were in turn selected for M from the automated procedure of configuration 4, even being

653

insensitive to river discharge. The error was also found in Pontelagoscuro images, but it was less severe due

654

to the different extent of the river, even if it can be responsible of the missing information for the greater

655

river discharge peaks, since, as mentioned above, those are related to very few pixels. To correct the issue,

656

several attempts were done using GEE “register” function without satisfying results, mainly because of the

657

presence of geometric distortions between the analyzed Sentinel-2 images. The correction of this effect

658

should be therefore entrusted to the data provider, in order to increase the multi-temporal consistency of

659

Sentinel-2 dataset.

660

The M pixels chosen over the fields, in the north of Figure 10c, highlight two additional limitations of the

661

automatic procedure. First, the vegetated pixels, which are poorly correlated with both C and W , are not

662

taken in account by the configuration 4, and can be potentially wrongly selected for M . Second, the number

663

of pixels sensitive to river discharge variation can change to not only rivers but also sections of the river

664

itself. The use of the percentile to obtain M pixels is needed for the portability of the methodology between

665 rivers of different sizes, but it could lead to select too many of them, especially for rivers with rectangular
 666 sections or narrow flooded area. In Montemolino case, for example, the most sensitive area is actually
 667 narrow and can be easily identified by looking at the M pixels selected by the calibrated procedures
 668 (configurations 1-3: green dot, blue area and red area, respectively, in Figure 10c): they fall in the same
 669 area inside the river, which is periodically submerged (it is shown in light blue color in Figure 10d: in some
 670 periods it dries, therefore the average reflectance and the standard deviation are higher with respect to
 671 the rest of the river area) and it is therefore sensitive to river discharge variation. No further sensitive areas
 672 can be clearly distinguished, probably due to the presence of the riparian vegetation, which obstruct their
 673 detection.

674 Once the C -, W - and M -mask were obtained, the C and W timeseries were used to estimate the river
 675 discharge with the CMW and the CM approaches, using both the uncalibrated and calibrated M timeseries.
 676 The performances of the obtained approaches for the river discharge estimation are shown in Table 2. The
 677 misregistration issue affects particularly the uncalibrated procedure, which showed the lowest
 678 performance in terms of R_s , R_p , $RMSE$ and NS . The calibrated methods were also affected, with score values
 679 worse than those obtained at Pontelagoscuro. Nevertheless, the relevance of considering sediment
 680 transport was again confirmed by the stable improvement in performances comparing the CMW and the
 681 CM results. Regarding the CMW approach, the configurations that show the best performance were the
 682 first two, where just one pixel and an area of 3×3 pixels were considered. This was due to the moderate
 683 width of the river, which requires a limited number of pixels to obtain the full river discharge variation.

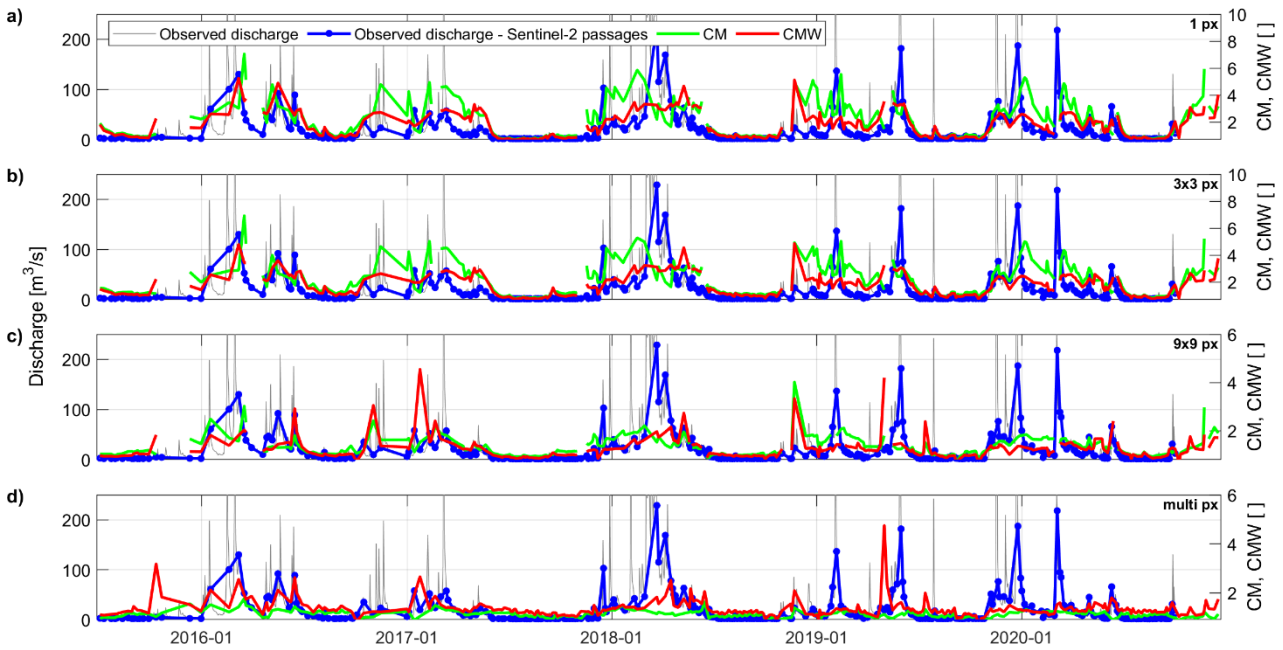
684 *Table 2: Performance scores for the different configurations (1 for single pixel, 2 for 3×3 pixels, 3 for 9×9 pixels and 4 for multiple*
 685 *pixels) for the period 2015-2020 at Montemolino station. Spearman (R_s) and Pearson (R_p) correlations are calculated between the*
 686 *CMW or CM timeseries and the ground observed river discharge. Root Mean Square Error ($RMSE$), relative Root Mean Square Error*
 687 *($rRMSE$) and Nash-Sutcliffe efficiency, NS , refer to the simulated discharge calculated according eq. 4 and the ground observed*
 688 *discharges.*

	CM	CMW
--	------	-------

Config.	R_s	R_p	$RMSE$	$rRMSE$	NS	R_s	R_p	$RMSE$	$rRMSE$	NS
	[-]	[-]	[m ³ /s]	[-]	[-]	[-]	[-]	[m ³ /s]	[-]	[-]
1	0.87	0.46	26.79	1.59	0.15	0.91	0.63	24.09	1.43	0.32
2	0.86	0.43	27.29	1.62	0.12	0.91	0.62	23.01	1.37	0.38
3	0.76	0.40	27.24	1.64	0.12	0.82	0.44	25.88	1.54	0.21
4	0.44	0.27	29.56	1.70	0.07	0.57	0.36	27.37	1.58	0.20

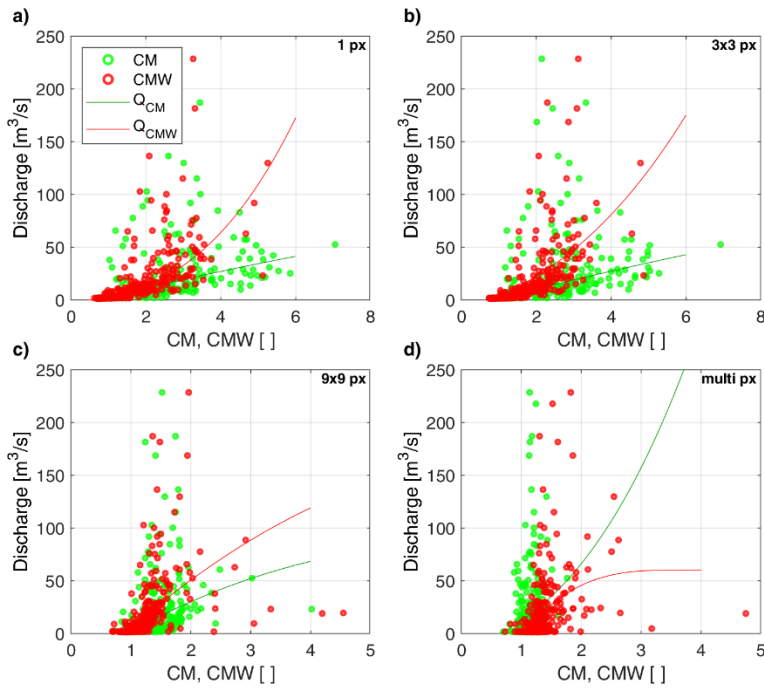
689

690 Figure 12 shows the timeseries of the ground observed discharge and those derived by *CMW* and *CM*
691 approaches at Montemolino station for all the configurations. The quality of the uncalibrated procedure
692 (Figure 12d) is relatively poor: even if flood periods were correctly detected, the surplus of *M* pixels and the
693 registration issue of Sentinel-2 data deteriorates the information. The calibrated procedure shows instead
694 better results (Figures 12a, b, c), even if all the river discharge peaks were underestimated, probably due
695 to saturation issues. Low flow conditions were quite well represented in most configurations. These results
696 are confirmed by the scatter plot shown in Figure 13: the configurations 1 and 2 (panel a and b) show a less
697 sparse couples with respect of the configuration 3 and 4 (panel c and d), which suffer from great noises.
698 This is also confirmed by the behavior of the fitting functions shown in the scatter plots, particularly for the
699 *CMW* ones (red lines, Figure 13): those in panel a and b show the steep behavior expected due to the
700 saturation of the *M* pixels, while the noises of configuration 3 and 4 (panel c and d) cause the obtained
701 relationships to display a milder slope.



702

703 *Figure 12: Tiber River at the Montemolino station: comparison in terms of temporal series between the ground observed discharge and the CM (green) or CMW (red) timeseries for the configuration 1 (single pixel, panel a), configuration 2 (3x3 averaged pixel, panel b), configuration 3 (9x9 averaged pixel, panel c) and for configuration 4 (multiple uncalibrated pixels, panel d). The observed discharge is shown in light grey for the full timeseries and in blue for the measurements coincident to Sentinel-2 overpass.*

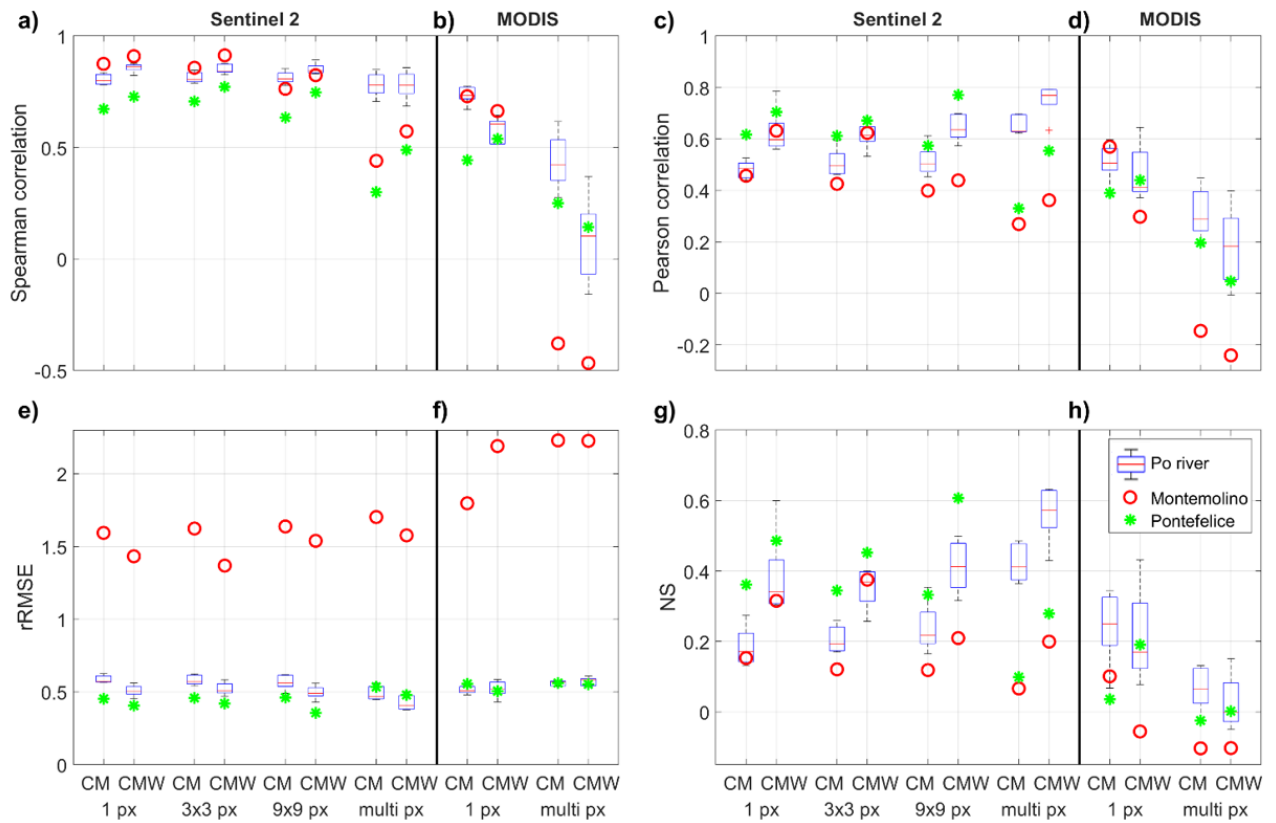


707

708 *Figure 63: Tiber River at the Montemolino station: comparison in terms of scatter plots between the ground observed discharge and the CM (green) or CMW (red) timeseries for configuration 1 (single pixel, panel a), configuration 2 (3x3 averaged pixel, panel b), configuration 3 (9x9 averaged pixel, panel c) and for configuration 4 (multiple uncalibrated pixels, panel d). The calibrated relationships of the estimated river discharge are shown as solid lines for both CM (dark green) and CMW (red).*

712 **4.4 Comparison between Sentinel-2 and MODIS: overall results assessment**

713 After the confirmation of the thresholds adopted and the acceptability of the results for both sites of
 714 Pontelagoscuro and Montemolino, the *CM* and *CMW* approaches were applied also to the remaining five
 715 sites by using data from both Sentinel-2 and MODIS. For Sentinel-2 all the configurations were considered,
 716 whereas with MODIS data only configurations 1 and 4 were adopted, because, in the configurations 2 and
 717 3, the final size of the resampled MODIS pixel is too large to represent accurately the river and to apply the
 718 approaches with reasonable results. The performance indices obtained between the *CM* and *CMW* ratios
 719 and the ground measurements of river discharge at all the seven sites are shown in Figure 14.



720
 721 *Figure 14: Spearman correlation, R_s (a, b), Pearson correlation R_p (c, d) between *CM* and *CMW* timeseries and ground observed*
 722 *discharges, relative Root Mean Square Error, $rRMSE$ (e, f) and Nash Sutcliffe efficiency, NS (g, h) between estimated and ground*
 723 *observed discharges by using Sentinel-2 (a, c, e, g) and MODIS (b, d, f, h) data as input for the different configurations. For the five*
 724 *sites of the Po River, the results are represented with box plots, whereas with stars and circles for the two stations of the Tiber River.*

725 The *CM* approach showed always lower or at most equal value of R_s or R_p compared to those obtained
 726 from the *CMW* approach when Sentinel-2 data were considered. Conversely, if MODIS data were adopted,
 727 *CM* had higher correlations than *CMW* (It should be noticed that the obtained results are lower than those

728 of Tarpanelli et al. (2013;2020) because of differences in the adopted methodology, e.g., cloud masking,
729 temporal interpolation). This was mainly due to the complexity in the selection of the W pixels in the latter
730 approach when MODIS data are adopted. In fact, because its spatial resolution was similar to or bigger than
731 the average width of the selected rivers, for low flows, the W pixels can be contaminated by a land fraction,
732 adding a source of noise to the obtained CMW index and causing a loss in performance with respect to the
733 original CM approach. Similarly, in the configuration 4, M is represented by the pixels that show low
734 correlations with both water and soil, but since W signal was contaminated, the selection of M pixels was
735 not optimal and consequently the performances were low. Moreover, the low spatial resolution of MODIS
736 causes the existence of many mixed pixels composed by fractions of terrain, water and vegetation, whose
737 signals will be poorly correlated with both C and W and thus erroneously selected for M . These
738 considerations were particularly exacerbated at the Tiber River sites, where the width of the river was
739 extensively lower than the MODIS resolution. Configuration 4 showed good results when Sentinel-2 data
740 were used, with R_s similar to those obtained for the other configurations whereas R_p higher for the Po
741 River. Slight worse performances were obtained for the Tiber River for both R_s and R_p . Specifically, the
742 performances obtained by configuration 4 at the Tiber River were affected by the presence of the riparian
743 vegetation and various agricultural fields in the neighboring of the river, together with the Sentinel-2 images
744 registration issue as detailed in the Section 4.3.

745 Many of the considerations made for the correlations are confirmed for the performances of NS and $rRMSE$
746 for the Po River. Instead, for the Tiber River the results at the two sites were very different each other.
747 Montemolino showed very low performances, particularly in terms of $rRMSE$, mainly due to the registration
748 issues of the Sentinel-2 images as discussed above. It is worth to notice that, despite the indices derived for
749 the configuration 4 did not take in account any in-situ data in the calibration of C , M and W location, the
750 knowledge of the ground observed discharge is still necessary to calibrate the parameters of eq. 4.
751 Notwithstanding this, the obtained results are promising and represent a step forward towards the
752 developing of a completely uncalibrated procedure to estimate river discharge from satellite data.

753 ***4.5 Discussion: potential and limitations***

754 The application of the CM approach to Sentinel-2 NIR data allows enlarging its field of application to
755 medium-small rivers, due to the higher spatial resolution of Sentinel-2 with respect to MODIS data, for
756 which the approach was developed. The analysis of very narrow rivers (10-20 meters) is still challenging
757 due to the presence of a co-registration issue among Sentinel-2 images. This issue has a strong impact on
758 the obtained performances (as observed at Montemolino station, section 4.3), since it causes many pixels
759 to vary “artificially” between soil and water, preventing the observation of the actual temporal variability
760 of the selected pixels reflectance signals. Moreover, it should be noticed that the application of the
761 sediment correction introduced in this study (CMW approach) to such narrow rivers should be avoided
762 when Sentinel-2 imagery are used. In fact, when the average river width is close to the size of a Sentinel-2
763 pixel, the probability that W pixels are contaminated by soil increases, thus adding a source of noise that
764 affects the obtained performances. The same effect has been observed from the application of the CMW
765 approach to MODIS data over the Po River (Section 4.4): the average Po River width at Pontelagoscuro
766 (~ 300 m) is close to the size of a MODIS pixel (~ 250 m), so the selected W pixels are contaminated by soil and
767 do not reflect the actual variability of the water reflectance. The sediment correction is therefore biased,
768 and the CMW approach is less performing than the original CM approach.

769 Notwithstanding these technological limitations, this study successfully demonstrated that high spatial
770 resolution and sediment correction allow to increase our ability to estimate river discharge from satellite
771 optical data. Moreover, the use of the uncalibrated approach to obtain M pixels locations is a first step
772 toward the application of the methodology on ungauged rivers. In order to achieve a fully uncalibrated
773 procedure, the approach needs further testing on multiple study areas worldwide, to assess the obtained
774 relationship and verify its applicability to different river regimes and areas. In addition, the role of
775 vegetation should be properly addressed, also to avoid vegetated pixels being wrongly selected for M , as
776 in the Montemolino station analysis (Figure 10c). Finally, a procedure to obtain the relationship between
777 river discharge and CM-based proxies without observed data has yet to be developed. However, the good
778 correlation between the obtained uncalibrated proxies and the observed river discharge indicates that their

779 information could be already exploited as qualitative signals, thus allowing the assessment of temporal
780 variability of river discharge (e.g., high flow and low flow periods) also on ungauged river systems.

781 Over gauged basins, estimated timeseries of river discharge can be obtained from CM indices using the
782 space-based rating curve concept. Indeed, the CM-based proxies are different from the satellite
783 measurements of hydraulic variables that exploit the same concept, such as altimeter river level data and
784 satellite measurements of river width. CM/CMW, in fact, do not estimate directly a physical quantity and
785 therefore they cannot be used directly in most available empiric or physical models. Nevertheless, also
786 satellite river level and river width data have inherent limitations. Altimeters are characterized by lower
787 temporal resolution and reduced spatial coverage with respect to optical sensors (Fekete et al., 2007;
788 Tarpanelli et al., 2021). On the other hand, optical data cannot be used to obtain information on the river
789 system in the presence of clouds. This issue is critical for river width estimation, which often lead the
790 researchers to discard many satellite images by imposing strict thresholds on the presence of clouds (Feng
791 et al., 2019), to consider cloud-free composite optical product with lower quality and temporal resolution
792 (Elmi et al., 2015; 2021) or to obtain river width measurements from sensors that are not affected by clouds,
793 such as microwave SAR (Synthetic Aperture Radar) sensors (Mengen et al., 2020). CM approaches, on the
794 other hand, show less sensitivity to presence of clouds, due to the use of a masking procedure of the pixels
795 affected by clouds and the averaging of large areas to obtain the values of the desired pixel values (e.g., in
796 this study, images with 70 % of cloudy pixels were processed, whereas in Feng et al. 2019 all images with
797 cloud percentage above 25 % were discarded). In addition, the techniques for estimating river width from
798 satellite are usually based on a water/non-water binary characterization (Elmi et al., 2015; 2021; Feng et
799 al., 2019; Mueller et al., 2016; Pavelsky et al., 2014; Sichangi et al., 2016), which limits the accuracy of the
800 obtained river width measurement to the pixel size of the relative satellite. Instead, CM approaches exploit
801 the measured reflectance values and should therefore theoretically allow for greater accuracy in river
802 discharge estimates with respect to river width analysis. Each of the existing methodologies for estimating
803 river discharge from satellite sensors has its own strengths and limitations, as river discharge is a rather
804 complex variable to observe because of its relationship with multiple physical quantities. Synergies

805 between the methodologies should be exploited to compensate their mutual weakness and maximize the
806 accuracy of the obtained product (Sichangi et al, 2016; Tarpanelli et al., 2017). Future research should be
807 oriented in this direction, to improve our ability to estimate river discharge from space.

808 **5. Conclusions**

809 Taking advantage of the high-resolution of Sentinel-2 data and their availability on Google Earth Engine
810 platform, a modification of the *CM* approach to estimate river discharge from NIR satellite data was
811 proposed. The method was enriched with the additional concept of exploiting a water (*W*) area to take in
812 account the variation of sediment load in the river. The averaging of multiple pixels to obtain *M* timeseries
813 was also proposed, to increase the methodology sensitivity to river discharge by avoiding as much as
814 possible the condition of completely flooded or dry *M*. Four different configurations were applied for the
815 selection of *M* pixels to test this concept, using as input Sentinel-2 data: three calibrated (1 pixel, 3x3 pixels
816 and 9x9 pixels) against the observed river discharge, and one uncalibrated, based on selecting as *M* pixels
817 those with low Spearman correlation with *W* or *C*. In order to test the advantages deriving from the use of
818 Sentinel-2 high-resolution data, the single pixel calibration and the uncalibrated one were also tested by
819 using as input the data from MODIS.

820 As a first step, a sensitivity analysis of each configuration thresholds was conducted at the station of
821 Pontelagoscuro, over the Po River: the thresholds related to the extraction of the *C*- and *W*-masks had little
822 incidence on the performance of the estimated river discharge and were fixed at the 5th percentile, while
823 the ones related to the selection of *M* generated greater variability, proving that further research is needed
824 to adapt the methodology to the different river regimes. The *CM* and *CMW* approaches were then applied
825 over seven stations from two rivers in Italy: at Pontelagoscuro, Sermide, Borgoforte, Cremona and Piacenza
826 along the Po River, and at Montemolino and Pontefelice along the Tiber River, for the period 2015-2020,
827 exploiting the capacities of Google Earth Engine cloud platform. Results indicated a stable performance
828 improvement of the *CMW* approach with respect to the *CM* approach for all the areas and configurations

829 when Sentinel-2 data were adopted, proving the relevance of taking in account the sediments to correctly
830 estimate the river discharge. The same advantages were not observed using MODIS data, due to the
831 impossibility of correctly account for the sediment contribution when the satellite pixel size is similar or
832 greater than the river width. Configuration 4 had proved the most effective at the Po River stations with
833 respect to the calibrated ones. Over the Tiber River, instead, the uncalibrated procedure failed to detect
834 the pixels sensitive to the variation of river discharge due to the presence of heavily vegetated pixels and
835 the misregistration issues of the Sentinel-2 images: the apparent movement of the full area led to the
836 selection of pixels that were not related to river discharge variation. The issue was present also at the Po
837 River stations, but to a lesser degree and with a less severe impact due to the greater river width. Two
838 additional sources of error were identified: the presence of cloud shadows and residual clouds not masked
839 by the cloud masking procedure, and the variability of the value of C , due to, e.g., the effect of soil moisture
840 variations or bidirectional reflectance effects.

841 This work represents a new step toward a non-calibrated methodology capable to obtain a river discharge
842 proxy over ungauged river, thanks to the development of an uncalibrated methodology to derive the
843 reflectance indices. The sediment contribution proved to be necessary to improve the river discharge
844 estimation. Still, well registered and frequent NIR high-resolution data are fundamental to increase the
845 accuracy of the retrievals. Further research is also needed to develop a relationship between the obtained
846 uncalibrated proxy and the river discharge that does not need calibration, based, e.g., on model data or
847 regionalization, in order to obtain a full uncalibrated methodology to estimate river discharge from space.
848 Moreover, the need of testing the procedure over different study areas and understanding the role of the
849 vegetation on river discharge estimation over highly vegetated cross sections should be also addressed, as
850 well as the development of the procedure to obtain a better threshold selection and greater sensitivity to
851 very high flows.

852 **Acknowledgment**

853 We would like to acknowledge Meteorological Environmental Earth Observation (MEEEO) company for their
854 role in download and subset Sentinel-2 images over Pontelagoscuro area. Also, we would like to acknowledge
855 the support of European Space Agency (ESA) through the project STREAMRIDE (Contract no. CCN1 to
856 4000126745/19/I-NB)

857

858 **References**

859 Abdalla, S., Abdeh Kolahchi, A., Adusumilli, S., Aich Bhowmick, S., Alou-Font, E., Amarouche, L., ... (2021)
860 Altimetry for the future: Building on 25 years of progress, *Advances in Space Research*. Doi:
861 <https://doi.org/10.1016/j.asr.2021.01.022>

862 Ahmed, R., Prowse, T., Dibike, Y., Bonsal, B., O'Neil, H. (2020). Recent Trends in Freshwater Influx to the Arctic
863 Ocean from Four Major Arctic-Draining Rivers. *Water* 12, 1189. Doi: <https://doi.org/10.3390/w12041189>

864 Ahn J. H. and Park, Y. J. (2020). Estimating Water Reflectance at Near-Infrared Wavelengths for Turbid Water
865 Atmospheric Correction: A Preliminary Study for GOCI-II. *Remote Sensing* 12(22), 3791. Doi:
866 <https://doi.org/10.3390/rs12223791>

867 Belloni, R., Camici, S., Tarpanelli, A. (2021). Towards the continuous monitoring of the extreme events
868 through satellite radar altimetry observations. *Journal of Hydrology*. In press. Doi:
869 <https://doi.org/10.1016/j.jhydrol.2021.126870>

870 Bjerklie D. M., Lawrence Dingman S., Vorosmarty C. J., Bolster C. H., Congalton R. G. (2003). Evaluating the
871 potential for measuring river discharge from space. *Journal of Hydrology* 278, 17–38. Doi:
872 [https://doi.org/10.1016/S0022-1694\(03\)00129-X](https://doi.org/10.1016/S0022-1694(03)00129-X)

873 Bjerklie D. M., Moller D., Smith L. C., Dingman S. L. (2005). Estimating discharge in rivers using remotely
874 sensed hydraulic information. *Journal of Hydrology* 309, 191–209. Doi:
875 <https://doi.org/10.1016/j.jhydrol.2004.11.022>

876 Boergens, E., Buhl, S., Dettmering, D., Klüppelberg, C., Seitz, F. (2017). Combination of multi-mission altimetry
877 data along the Mekong River with spatio-temporal kriging. *Journal of Geodesy* 91(5), 519-534. Doi:
878 <https://doi.org/10.1007/s00190-016-0980-z>

879 Brakenridge, G. R., Nghiem, S. V., Anderson, E., Mic, R (2007). Orbital microwave measurement of river
880 discharge and ice status. *Water Resources Research* 43, W04405. Doi: 10.1029/2006WR005238

881 Chandanpurkar H. A., Reager J. T., Famiglietti J. S., Syed T. H. (2017). Satellite- and Reanalysis-Based Mass
882 Balance Estimates of Global Continental Discharge (1993–2015). *Journal of Climate* 30(21), 8481–8495. Doi:
883 <https://doi.org/10.1175/JCLI-D-16-0708.1>

884 Crochemore, L., Isberg, K., Pimentel, R., Pineda, L., Hasan, A., Arheimer, B. (2020). Lessons learnt from
885 checking the quality of openly accessible river flow data worldwide. *Hydrological Sciences Journal* 65(5), 699-
886 711. Doi: <https://doi.org/10.1080/02626667.2019.1659509>

887 De Frasson R. P. M., Pavelsky T. M., Fonstad M. A., Durand M. T., Allen G. H., Schumann G., Lion C., Beighley
888 R. E., Yang X. (2019) Global Relationships Between River Width, Slope, Catchment Area, Meander
889 Wavelength, Sinuosity, and Discharge. *Geophysical Research Letters* 46(6), 3252–3262. Doi:
890 <https://doi.org/10.1029/2019GL082027>

891 Domeneghetti, A., Tarpanelli, A., Grimaldi, L., Brath, A., Schumann, G. (2018). Flow Duration Curve from
892 Satellite: Potential of a Lifetime SWOT Mission. *Remote Sensing* 10(7), 1107. Doi:
893 <https://doi.org/10.3390/rs10071107>

894 Durand M., Neal J., Rodríguez E., Andreadis K. M., Smith L. C., Yoon Y. (2014). Estimating reach-averaged
895 discharge for the River Severn from measurements of river water surface elevation and slope. *Journal of*
896 *Hydrology* 511, 92–104. Doi: <https://doi.org/10.1016/j.jhydrol.2013.12.050>

897 Dziubanski D. J, Franz K. J. (2016). Assimilation of AMSR-E snow water equivalent data in a spatially-lumped
898 snow model. *Journal of Hydrology* 540, 26–39. Doi: <https://doi.org/10.1016/j.jhydrol.2016.05.046>

899 Elmi O., Tourian M. J., Sneeuw N. (2015). River discharge estimation using channel width from satellite
900 imagery. *IEEE International Geoscience and Remote Sensing Symposium (IGARSS)*, 727-730. Doi:
901 <https://doi.org/10.1109/IGARSS.2015.7325867>.

902 Elmi O., Tourian M. J., Bárdossy A., Sneeuw N. (2021). Spaceborne River Discharge From a Nonparametric
903 Stochastic Quantile Mapping Function, *Water Resources Research* 57(12), e2021WR030277. Doi:
904 <https://doi.org/10.1029/2021WR030277>

905 Emery C. M., Paris A., Biancamaria S., Boone A., Calmant S., Garambois P.-A., Santos da Silva J. (2018). Large-
906 scale hydrological model river storage and discharge correction using a satellite altimetry-based discharge
907 product. *Hydrology and Earth System Sciences* 22, 2135–2162. Doi: [https://doi.org/10.5194/hess-22-2135-](https://doi.org/10.5194/hess-22-2135-2018)
908 2018

909 Fekete, B. and Vörösmarty, C. (2007). The current status of global river discharge monitoring and potential
910 new technologies complementing traditional discharge measurements. *Proceedings of the Predictions in*
911 *Ungauged Basins: PUB Kick-off (Proceedings of the PUB Kick-off meeting held in Brasilia, 20–22 November*
912 *2002)*. IAHS Publ. 309, 2007.

913 Fekete, B. M., Robarts, R. D., Kumagai, M., Nachtnebel, H. P., Odada, E., Zhulidov, A. V. (2015). Time for in-
914 situ renaissance. *Science* 349, 685–686. Doi: <https://doi.org/10.1126/science.aac7358>.

915 Feng D., Gleason C. J., Yang X., Pavelsky T. M. (2019). Comparing discharge estimates made via the BAM
916 algorithm in high-order Arctic rivers derived solely from optical CubeSat, Landsat, and Sentinel-2 data, *Water*
917 *Resources Research* 55(9), 7753-7771. Doi: <https://doi.org/10.1029/2019wr025599>

918 Garambois P.-A., Monnier J. (2015). Inference of effective river properties from remotely sensed observations
919 of water surface. *Advances in Water Resources* 79, 103–120. Doi:
920 <https://doi.org/10.1016/j.advwatres.2015.02.007>

921 Gentile, F., Bisantino, T., Corbino, R., Milillo, F., Romano, G., Liuzzi, G. T. (2010). Monitoring and analysis of
922 suspended sediment transport dynamics in the Carapelle torrent (southern Italy). *Catena* 80(1), 1–8. Doi:
923 <https://doi.org/10.1016/j.catena.2009.08.004>

924 Gilvear, D., Hunter, P., Higgins, T. (2007). An experimental approach to the measurement of the effects of
925 water depth and substrate on optical and near infra-red reflectance: A fieldbased assessment of the
926 feasibility of mapping submerged instream habitat. *International Journal of Remote Sensing*. *International*
927 *Journal of Remote Sensing*, 28(10), 2241-2256. Doi: <https://doi.org/10.1080/01431160600976079>.

928 Gleason C. J., Smith L. C. (2014). Toward global mapping of river discharge using satellite images and at-many-
929 stations hydraulic geometry. *Proceedings of the National Academy of Sciences* 111(13), 4788–4791. Doi:
930 <https://doi.org/10.1073/pnas.1317606111>

931 Gleason C. J., Garambois P. A., Durand M. (2017). Tracking River Flows from Space. *Eos Transactions American*
932 *Geophysical Union* 99(1). Doi: <https://doi.org/10.1029/2017EO078085>

933 Gleason C. J., Durand M. T. (2020). Remote Sensing of River Discharge: A Review and a Framing for the
934 Discipline. *Remote Sensing*, 12, 1107. Doi: <https://doi.org/10.3390/rs12071107>

935 Gorelick, N., Hancher, M., Dixon, M., Ilyushchenko, S., Thau, D., Moore, R. (2017). Google Earth Engine:
936 Planetary-scale geospatial analysis for everyone. *Remote sensing of Environment* 202, 18-27. Doi:
937 <https://doi.org/10.1016/j.rse.2017.06.031>

938 Hannah, D. M., Demuth, S., Van Lanen, H. A. J., Looser, U., Prudhomme, C., Rees, G., Stahl, K., Tallaksen, L.
939 M. (2011). Large scale river flow archives: importance, current status and future needs. *Hydrological*
940 *Processes* 25(7), 1191-1200. Doi: <https://doi.org/10.1002/hyp.7794>

941 Huo J., Qu X., Zhu D., Yuan Z., Zeng Z. (2021). Runoff monitoring in the Lhasa River Basin using passive
942 microwave data, *International Journal of Applied Earth Observation and Geoinformation* 103, 102486. Doi:
943 <https://doi.org/10.1016/j.jag.2021.102486>

944 Jodar J., Carpintero E., Martos-Rosillo S., Ruiz-Constan A., Marin-Lechado C., Cabrera-Arrabal J. A., Navarrete-
945 Mazariegos E., Gonzalez-Ramon A., Lamban L. J., Herrera C., González-Dugo M. P. (2018). Combination of
946 lumped hydrological and remote-sensing models to evaluate water resources in a semi-arid high altitude
947 ungauged watershed of Sierra Nevada (Southern Spain). *Science of Total Environment* 625, 285–300. Doi:
948 <https://doi.org/10.1016/j.scitotenv.2017.12.300>

949 Keesstra, S. D., Davis, J., Masselink, R. H., Casalí, J., Peeters, E. T., Dijkma, R. (2019). Coupling hysteresis
950 analysis with sediment and hydrological connectivity in three agricultural catchments in Navarre, Spain.
951 *Journal of Soils and Sediments* 19(3), 1598-1612. Doi: <https://doi.org/10.1007/s11368-018-02223-0>

952 Kremezi, M. and Karathanassi, V. (2019). Correcting the BRDF effects on Sentinel-2 ocean images. *Proc. SPIE*
953 11174, *Seventh International Conference on Remote Sensing and Geoinformation of the Environment*
954 (RSCy2019), 111741C (27 June 2019). Doi: <https://doi.org/10.1117/12.2533653>

955 Larnier K., Monnier J., Garambois P.-A., Verley J. (2019). River Discharge and Bathymetry Estimations from
956 SWOT Altimetry Measurements. *Inverse Problems in Science and Engineering* 29(6), 1-31. Doi:
957 <https://doi.org/10.1080/17415977.2020.1803858>

958 Lin P., Pan M., Beck H. E., Yang Y., Yamazaki D., Frasson R., David C. H., Durand M., Pavelsky T. M., Allen G.
959 H., Gleason C. J., Wood E. F. (2019). Global Reconstruction of Naturalized River Flows at 2.94 Million Reaches.
960 *Water Resources Research* 55(8), 6499–6516. Doi: <https://doi.org/10.1029/2019WR025287>

961 Lopez P. L., Sutanudjaja E. H., Schellekens J., Sterk G., Bierkens M. F. P. (2017). Calibration of a large-scale
962 hydrological model using satellite-based soil moisture and evapotranspiration products. *Hydrology and Earth
963 System Sciences* 21, 3125–3144. Doi: <https://doi.org/10.5194/hess-21-3125-2017>

964 Malinowski, R., Groom, G., Schwanghart, W., Goswin, H. (2015). Detection and Delineation of Localized
965 Flooding from WorldView-2 Multispectral Data. *Remote Sensing* 7, 14853-14875. Doi:
966 <https://doi.org/10.3390/rs71114853>.

967 Mengen D., Ottinger M., Leinenkugel P., Ribbe L. (2020). Modeling River Discharge Using Automated River
968 Width Measurements Derived from Sentinel-1 Time Series. *Remote Sensing* 12(19), 3236. Doi:
969 <https://doi.org/10.3390/rs12193236>

970 Mueller N., Lewis A., Roberts D., Ring S., Melrose R., Sixsmith J., Lymburner L., McIntyre A., Tan P., Curnow
971 S., Ip A. (2016). Water observations from space: Mapping surface water from 25 years of Landsat imagery
972 across Australia, *Remote Sensing of Environment* 174, 341-352. Doi:
973 <https://doi.org/10.1016/j.rse.2015.11.003>

974 Nash, J. E. and Sutcliffe, J. V. (1970). River Flow Forecasting through Conceptual Model. Part 1 - A Discussion
975 of Principles. *Journal of Hydrology*, 10, 282-290. Doi: [http://dx.doi.org/10.1016/0022-1694\(70\)90255-6](http://dx.doi.org/10.1016/0022-1694(70)90255-6)

976 Neal J., Schumann G., Bates P., Buytaert W., Matgen P., Pappenberger F. (2009). A data assimilation approach
977 to discharge estimation from space. *Hydrological Processes* 23(25), 3641–3649. Doi:
978 <https://doi.org/10.1002/hyp.7518>

979 Nohara, D., Kitoh, A., Hosaka, M., Oki, T. (2006). Impact of Climate Change on River Discharge Projected by
980 Multimodel Ensemble. *Journal of Hydrometeorology* 7(5), 1076-1089. Doi:
981 <https://doi.org/10.1175/JHM531.1>

982 Oubanas H., Gejadze I., Malaterre P.-O., Mercier F. (2018). River discharge estimation from synthetic SWOT-
983 type observations using variational data assimilation and the full Saint-Venant hydraulic model. *Journal of*
984 *Hydrology* 559, 638–647. Doi: <https://doi.org/10.1016/j.jhydrol.2018.02.004>

985 Paris, A., Dias de Paiva, R., Santos da Silva, J., Medeiros Moreira, D., Calmant, S., Garambois, P. A., Collischonn,
986 W., Bonnet, M. P., Seyler, F. (2016). Stage-discharge rating curves based on satellite altimetry and modeled
987 discharge in the Amazon basin. *Water Resources Research* 52(5), 3787-3814. Doi:
988 <https://doi.org/10.1002/2014WR016618>

989 Parr D., Wang G., Bjerklie D. (2015). Integrating Remote Sensing Data on Evapotranspiration and Leaf Area
990 Index with Hydrological Modeling: Impacts on Model Performance and Future Predictions. *Journal of*
991 *Hydrometeorology* 16(5), 2086–2100. Doi: <https://doi.org/10.1175/JHM-D-15-0009.1>

992 Pavelsky T. M. (2014). Using width-based rating curves from spatially discontinuous satellite imagery to
993 monitor river discharge, *Hydrological Processes* 28, 3035-3040. Doi: <https://doi.org/10.1002/hyp.10157>

994 Pekel, J. F., Cottam, A., Gorelick, N., Belward, A. S. (2016). High-resolution mapping of global surface water
995 and its long-term changes. *Nature* 540, 418–422. Doi: <https://doi.org/10.1038/nature20584>

996 Piecuch, C. G., Bittermann, K., Kemp, A. C., Ponte, R. M., Little, C. M., Engelhart, S. E., Lentz, S. J. (2018). River-
997 discharge effects on United States Atlantic and Gulf coast sea-level changes. *Proceedings of the National*
998 *Academy of Sciences* 115(30), 7729-7734. Doi: <https://doi.org/10.1073/pnas.1805428115>

999 Schwatke, C., Dettmering, D., Bosch, W., Seitz, F. (2015). DAHITI—an innovative approach for estimating water
1000 level time series over inland waters using multi-mission satellite altimetry. *Hydrology and Earth System*
1001 *Sciences* 19(10), 4345-4364. Doi: <https://doi.org/10.5194/hess-19-4345-2015>

1002 Shi, Z., Chen, Y., Liu, Q., Huang, C. (2020). Discharge Estimation Using Harmonized Landsat and Sentinel-2
1003 Product: Case Studies in the Murray Darling Basin. *Remote Sensing* 12(17), 2810. Doi:
1004 <https://doi.org/10.3390/rs12172810>

1005 Sichangi A. W., Wang L., Yang K., Chen D., Wang Z., Li X., Zhou J., Liu W., Kuria D. (2016). Estimating continental
1006 river basin discharges using multiple remote sensing data sets, *Remote Sensing of Environment* 179, 36-53.
1007 Doi: <https://doi.org/10.1016/j.rse.2016.03.019>

1008 Syed T. H., Famiglietti J. S., Chen J., Rodell M., Seneviratne S. I., Viterbo P., Wilson C. R. (2005). Total basin
1009 discharge for the Amazon and Mississippi River basins from GRACE and a land-atmosphere water balance.
1010 *Geophysical Research Letters* 32, L24404. Doi: <https://doi.org/10.1029/2005GL024851>

1011 Tarpanelli, A., Brocca, L., Lacava, T., Melone, F., Moramarco, T., Faruolo, M., Pergola, N., Tramutoli, V. (2013).
1012 Toward the estimation of river discharge variations using MODIS data in ungauged basins. *Remote Sensing*
1013 *of Environment* 136, 47–55. Doi: <https://doi.org/10.1016/j.rse.2013.04.010>

1014 Tarpanelli, A., Amarnath, G., Brocca, L., Massari, C., Moramarco, T. (2017). Discharge estimation and
1015 forecasting by MODIS and altimetry data in Niger-Benue River. *Remote Sensing of Environment* 195, 96–106.
1016 Doi: <https://doi.org/10.1016/j.rse.2017.04.015>

1017 Tarpanelli, A., Iodice, F., Brocca, L., Restano, M., Benveniste, J. (2020). River Flow Monitoring by Sentinel-3
1018 OLCI and MODIS: Comparison and Combination. *Remote Sensing* 12, 3867. Doi:
1019 <https://doi.org/10.3390/rs12233867>

1020 Tarpanelli, A., Camici, S., Nielsen, K., Brocca, L., Moramarco, T., Benveniste, J. (2021). Potentials and
1021 limitations of Sentinel-3 for river discharge assessment. *Advances in Space Research* 68(2), 593-606. Doi:
1022 <https://doi.org/10.1016/j.asr.2019.08.005>

1023 Tourian, M. J., Sneeuw, N., Bardossy, A. (2013). A quantile function approach to discharge estimation from
1024 satellite altimetry (ENVISAT). *Water Resources Research* 49(7), 4174–4186. Doi:
1025 <https://doi.org/10.1002/wrcr.20348>.

1026 Tourian, M. J., Tarpanelli, A., Elmi, O., Qin, T., Brocca, L., Moramarco, T., Sneeuw, N. (2016). Spatiotemporal
1027 densification of river water level time series by multimission satellite altimetry. *Water Resources Research*
1028 52, 1140-1159. Doi: <https://doi.org/10.1002/2015WR017654>.

1029 Tourian, M. J., Schwatke, C., Sneeuw, N. (2017). River discharge estimation at daily resolution from satellite
1030 altimetry over an entire river basin. *Journal of Hydrology* 546, 230-247. Doi:
1031 <https://doi.org/10.1016/j.jhydrol.2017.01.009>.

1032 Vörösmarty, C., Askew, A., Grabs, W., Barry, R. G., Birkett, C., Döll, P., Goodison, B., Hall, A., Jenne, R., Kitaev,
1033 L., Landwehr, J., Keeler, M., Leavesley, G., Schaake, J., Strzepek, K., Sundarvel, S. S., Takeuchi, K., Webster, F.
1034 (2001). Global water data: A newly endangered species. *EOS Transactions* 82, 54-54. Doi:
1035 <https://doi.org/10.1029/01EO00031>

1036 Wagner, W., Lemoine, G., Rott, H. (1999). A method for estimating soil moisture from ERS scatterometer and
1037 soil data. *Remote Sensing of Environment* 70, 191–207. Doi: [https://doi.org/10.1016/S0034-4257\(99\)00036-](https://doi.org/10.1016/S0034-4257(99)00036-)
1038 X

1039 Wulf H., Bookhagen B., Scherler D. (2016). Differentiating between rain, snow, and glacier contributions to
1040 river discharge in the western Himalaya using remote-sensing data and distributed hydrological modeling.
1041 *Advances in Water Resources* 88, 152–169. Doi: <https://doi.org/10.1016/j.advwatres.2015.12.004>

1042 Yang J., Huang X., Tang Q. (2020). Satellite-derived river width and its spatiotemporal patterns in China during
1043 1990–2015, *Remote Sensing of Environment* 247, 111918. Doi: <https://doi.org/10.1016/j.rse.2020.111918>

1044 Zakharova E., Nielsen K., Kamenev G., Kouraev A. (2020). River discharge estimation from radar altimetry:
1045 Assessment of satellite performance, river scales and methods, *Journal of Hydrology* 583, 124561. Doi:
1046 <https://doi.org/10.1016/j.jhydrol.2020.124561>

1047 Zhang Y., Pan M., Wood E. F. (2016). On Creating Global Gridded Terrestrial Water Budget Estimates from
1048 Satellite Remote Sensing. *Surveys in Geophysics* 37, 249–268. Doi: <https://doi.org/10.1007/s10712-015->
1049 9354-y

1050

1051